

1 Supplementary material

2 Biotic interactions may be more important 3 than climate for shaping animal niches

4

5

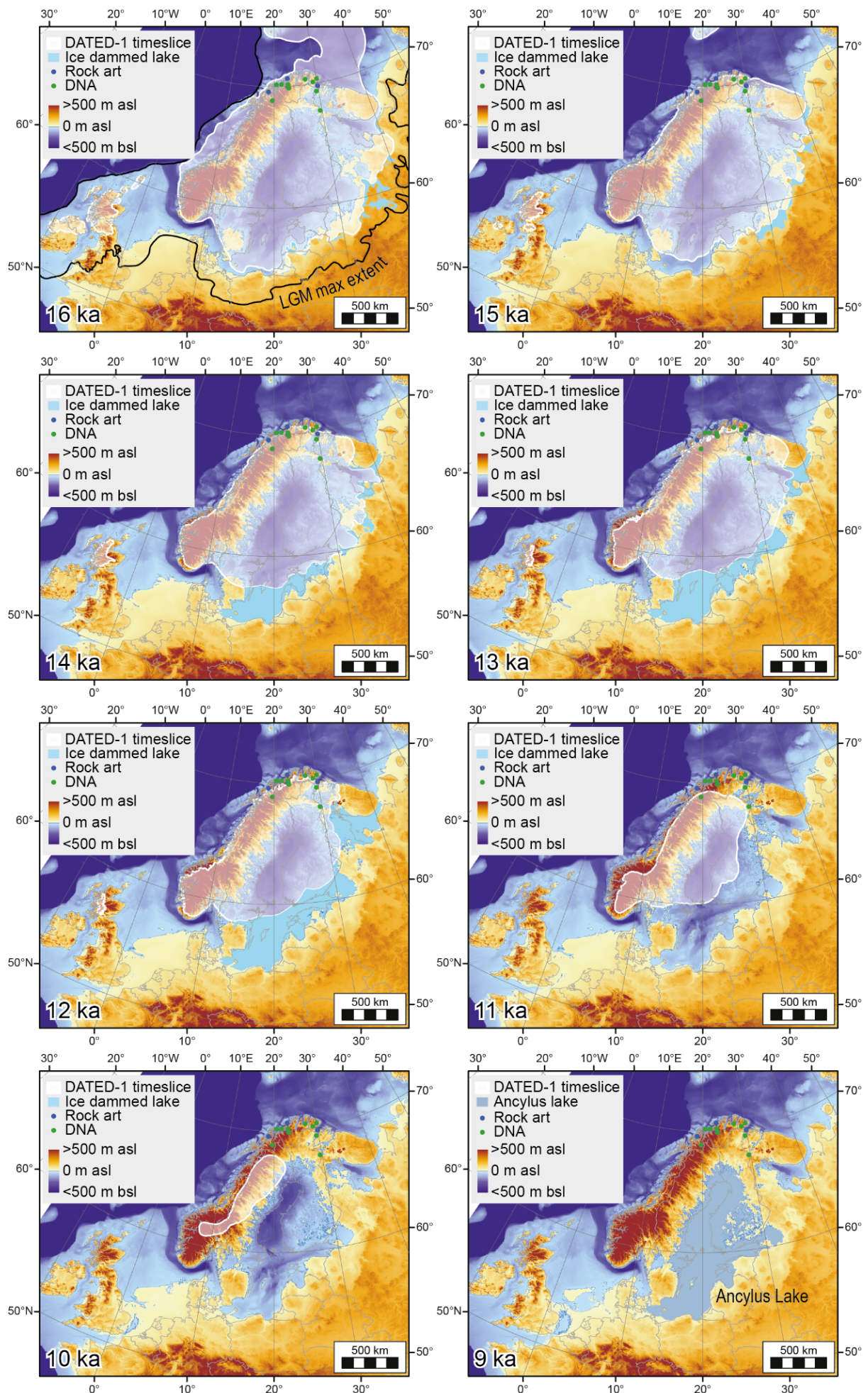
6	1. Changing landscapes during deglaciation	1
7	2. Catchments and taphonomy	4
8	2.1 The theory	4
9	2.2 Our Application	5
10	3. Mitigating false positives from sampling to final dataset	11
11	4. The full animal sedaDNA record	15
12	5. Comparison of bone and sedaDNA records	16
13	6. Animal frequency over time	17
14	7. Plant-herbivore co-occurrence networks	20
15	7.1 Reindeer co-occurrence	20
16	7.2 Grey-sided vole co-occurrence	20
17	7.3 Norwegian lemming co-occurrence	20
18	7.4 Willow ptarmigan	21
19	7.5 Strength of plant-herbivore associations	23
20	8. Niche change for individual species	24
21	8.1 Changes in niche sizes	24
22	8.2 Reindeer niche	25
23	8.3 Grey-sided vole niche	29
24	8.4 Norwegian lemming niche	33
25	8.5 Willow ptarmigan niche	37
26	8.6 Frog niche	40
27	9 Change in environmental space	44
28	10 Niche overlap among key herbivores	47
29	11 Lemming and tundra vole 16S reference sequences	57
30	Supplementary references	59

31

32

33 1. Changing landscapes during deglaciation

34 During deglaciation, there were changes in glacial ice extent ^{1,2}, sea level ^{3,4}, and the size of
35 ice-dammed lakes ⁵ in northern Fennoscandia. When these factors are combined, it
36 becomes clear how the landscape formed a mosaic of dispersal barriers and highways for
37 different organisms. For some species, like geese and freshwater plants and invertebrates,
38 the large ice-dammed lakes that temporarily formed during ice melt may have represented
39 dispersal highways while they at the same time represented dispersal barriers to terrestrial
40 mammals.



42

43 *Supplementary Fig. 1.1 sedaDNA coring sites and rock art sites in relation to the*
44 *glacio-isostatically modelled palaeo-topography⁶, most-credible ice-sheet extent¹, and*
45 *predicted paleo-shorelines determining the extent and position of ice-dammed lakes and*
46 *marine sectors, for one thousand year time slices between 16 and 9 ka.*

47

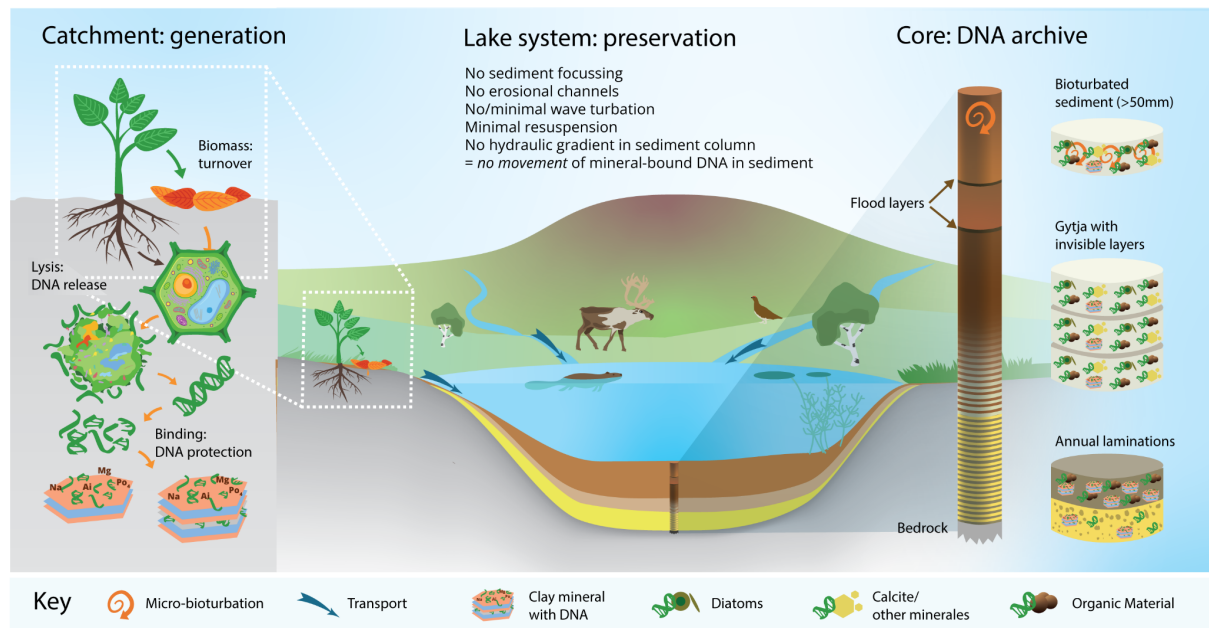
48 2. Catchments and taphonomy

49 2.1 The theory

50 The fate of DNA from deposition to sediments is shown in Fig. S2.1. The source of plant,
51 mammal and much invertebrate sedaDNA is the soil and slopes of the hydrological
52 catchment of the lake⁷⁻¹⁰. After cell lysis during decomposition of litter and below ground
53 biomass (e.g. roots), DNA is released into the pore waters of the soil¹¹. This DNA has a
54 half-life in aqueous solution of minutes to a few hours if not protected from UV light,
55 temperature and nucleases in the microbial environment (Collins et al. 2018). However, a
56 small proportion is protected from further fragmentation by being adsorbed onto minerals,
57 particularly clays, and organic complexes^{12,13}. Some clays can even incorporate DNA
58 strands into their structure (nanoconfinement) and be preserved for two million years¹⁴.
59 These mineral and organic particles are then transported by overland water flow with their
60 adsorbed DNA to streams and rivers and then into lakes, estuaries or coastal areas^{7,15}.
61 Because the DNA is attached preferentially to the finer part of the suspended river load it is
62 carried out into the lake and settles out with decreasing thickness away from the river input.
63 In small-medium sized lakes of shallow to moderate depth (4-30m) this fine layer is
64 minimally disturbed although micro-scale bioturbation can be caused by seasonal
65 resuspension, wave action (down to c. 4m) and micro-bioturbation by benthic fauna such as
66 tubificid worms. This will cause mm to a few cm of mixing and depending upon the
67 accumulation rate may blur the resultant age model by a few years. Where this does not
68 occur, annual laminations (pseudo-varves) can be preserved along with flood layers caused
69 by larger floods, which are routinely used to create flood chronologies¹⁶. This has been
70 validated by sedaDNA which reflects a larger source area within the catchment than
71 between the floods reflecting the expansion of 'contributing area' of sediment for larger
72 hydrological events¹⁷. Thousands of studies of lake sediment fabric (3D arrangement of
73 particles) have revealed that the movement of clay and fine organic matter does not occur in
74 lakes sediment columns due to saturation of pore water spaces and a lack of hydraulic
75 gradient. Lake sediments can be disturbed where bed slopes are steep causing sediment
76 focusing and if currents or channels erode into the lake floor producing problems particularly
77 on larger and more bathymetrically complex lakes. However, in small shallow single basin
78 lakes, the sediment column preserves a faithful record of changes over time in sediment
79 mineralogy, geochemistry, microfossils and organic molecules bound to mineral and organic
80 particles with no vertical mixing. This can even be validated from sedaDNA which record to
81 within a few years changes in catchment vegetation or known introductions^{18,19}.

82

83



84

85 *Supplementary Fig 2.1. Catchment sources of DNA and its deposition in lake sediments. Left*
 86 *panel: Upon cell lysis, DNA is released and a small fraction bound to organic and mineral*
 87 *particles, especially clays. Central panel: Particles with DNA are transported by overland flow*
 88 *and streams to be deposited on lake floors. Right panel: The sediment column showing the*
 89 *sequential deposition of DNA, here shown as a transition from clastic dominated varved*
 90 *(annual laminated) to organic dominated (gyttja) sediments with flood layers.*
 91 *Microbioturbation in top sediment may cause mixing of a few centimetre. By selecting*
 92 *morphologically simple lakes where there is no sediment erosion and re-deposition, the DNA*
 93 *is stratified in a reliable time sequence.*

94

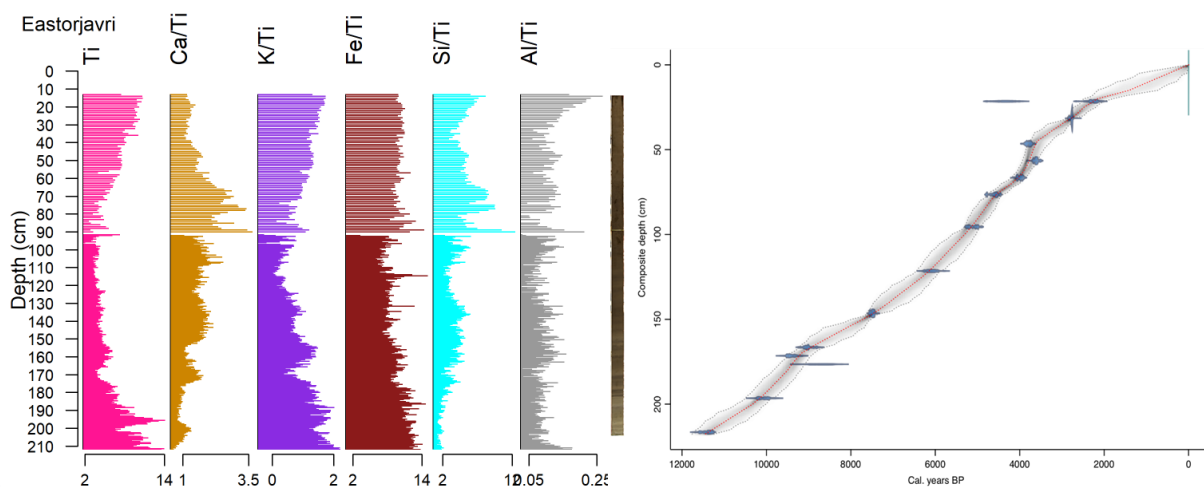
95 2.2 Our application

96 The catchments used in this study were chosen specifically so that taphonomic variation and
 97 complicating factors would be minimized following both our recent taphonomic studies⁸ and
 98 work elsewhere^{7,20}. So all our lakes were small (Table S1) being <60 ha (and mostly under
 99 15 ha) and shallow (under 15 m except for one), with simple bathymetry and gentle sediment
 100 surface slopes, negating any sediment-focussing effects. The lake catchments were also
 101 small with all except one being under 11 km². Additionally, most of the lakes were also in
 102 headwater catchments with no upstream lake inputs and all except one (Sandfjordalen) had
 103 no significant upstream floodplains or areas of colluvial sediment storage. That we do not
 104 have eroded sedaDNA is supported by a lack of any out-of sequence taxa or indeed taxa
 105 from previous warm-periods. The lakes are also dominated by in-lake organic production
 106 with relatively stable levels of mineral input except in the pre-Holocene (see XRF,
 107 Supplementary Fig. S2.2; For Kuutsjävi see²¹). Whilst Al is present in all cores, reflecting
 108 clays, it varies being low in sections with marine sands in Langfjordvannet and high in
 109 periods of fine marine sedimentation in Nordivatnet, whereas it is intermediate in all
 110 lacustrine sections. These results suggest adequate clay content in all sediments for the
 111 unbiased preservation of sedaDNA. Only two lakes have significant change in carbonate
 112 component in the early Holocene (Fig S2.2) and virtually no mineral in-wash bands were

113 seen, although the Si and Ti record of Langfjordvannet reflects sand inwashes. Otherwise, a
 114 lack of inwashes of silt-clay-organic matter and reworking is also supported by under 10 ^{14}C
 115 date reversals, with only 5 being older dates (i.e. possible unwashed organic matter) out of
 116 over 120 dates in the Holocene ²², with 5 of these coming from two sites (Sandfjordalen and
 117 Sierrevatnet, Supplementary Fig. 2.2). This is unlike the records from lakes in most other
 118 areas of Europe ⁷ and reflects the lack of agriculture in these catchments ²³. The result is that
 119 the input of *sedaDNA* to the lake was by continuous stream and lake-edge transfer and so
 120 would have reflected the presence of the taxa rather than a delayed response due to
 121 temporal sediment dynamics. Lastly, the lakes all share the same or similar geologies
 122 (metamorphic gneiss, schists and slates) which produce the same suit of clay minerals that
 123 include some swelling clays which have a high affinity to *sedaDNA* ²⁴. thus, the cores can be
 124 regarded as cumulative records as there is abundant evidence now that *sedaDNA* cannot be
 125 leached from one level to another in saturated lake sediments ¹⁸ – as also applies to other
 126 biomarkers and geochemistry. In this sedimentary context, DNA damage patterns are not
 127 required for ancient DNA authentication, and *sedaDNA* metabarcoding data are regarded as
 128 reliable source of chronological ecosystem changes ^{25–27}.

129

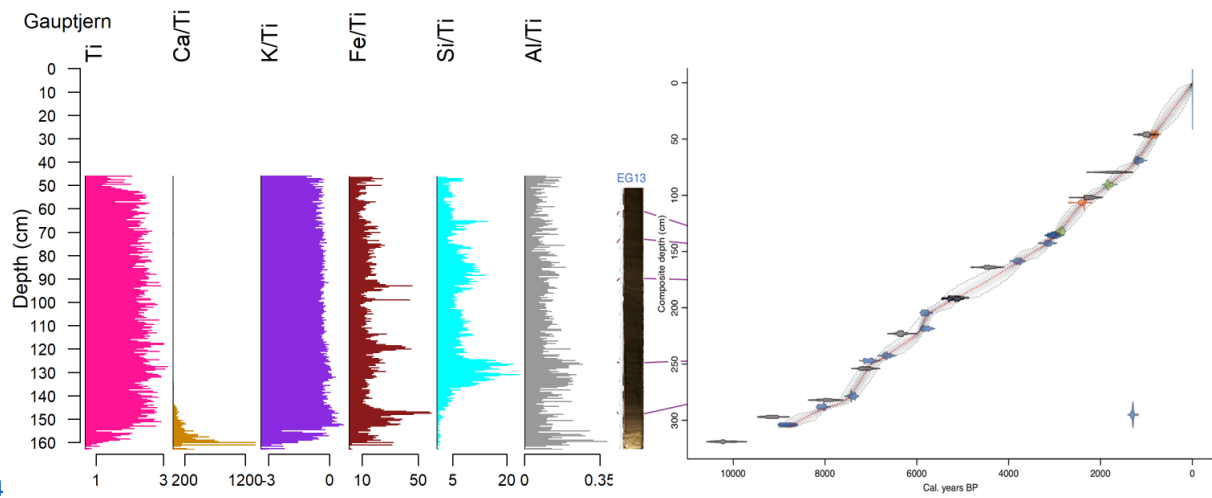
130



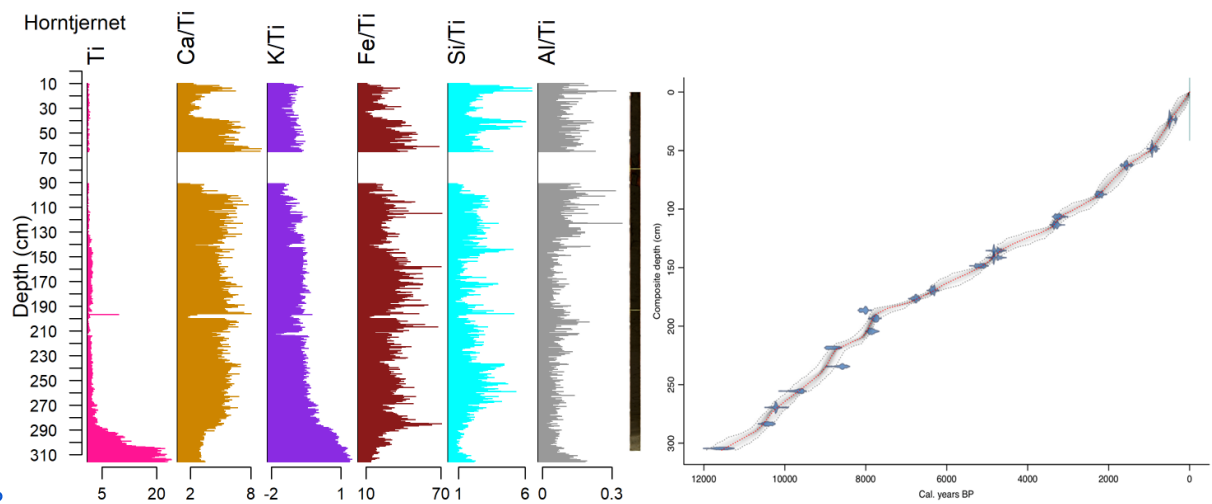
131

132

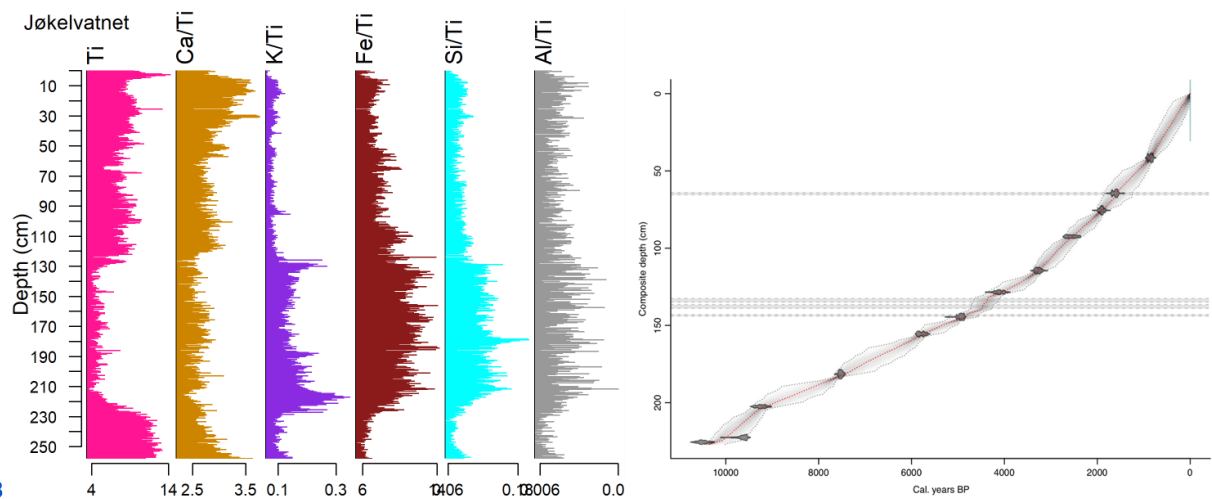
133



134

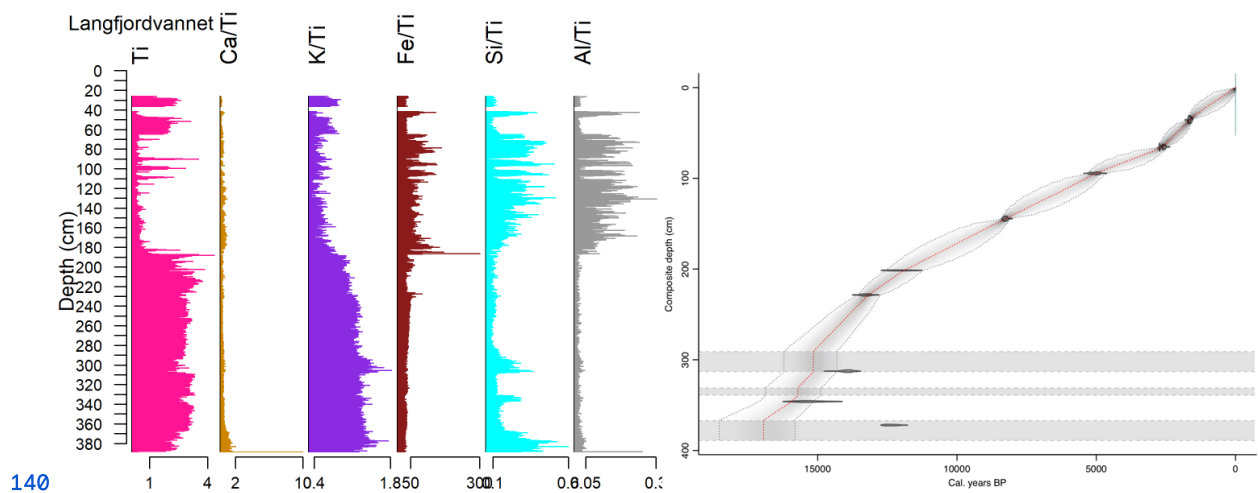


136



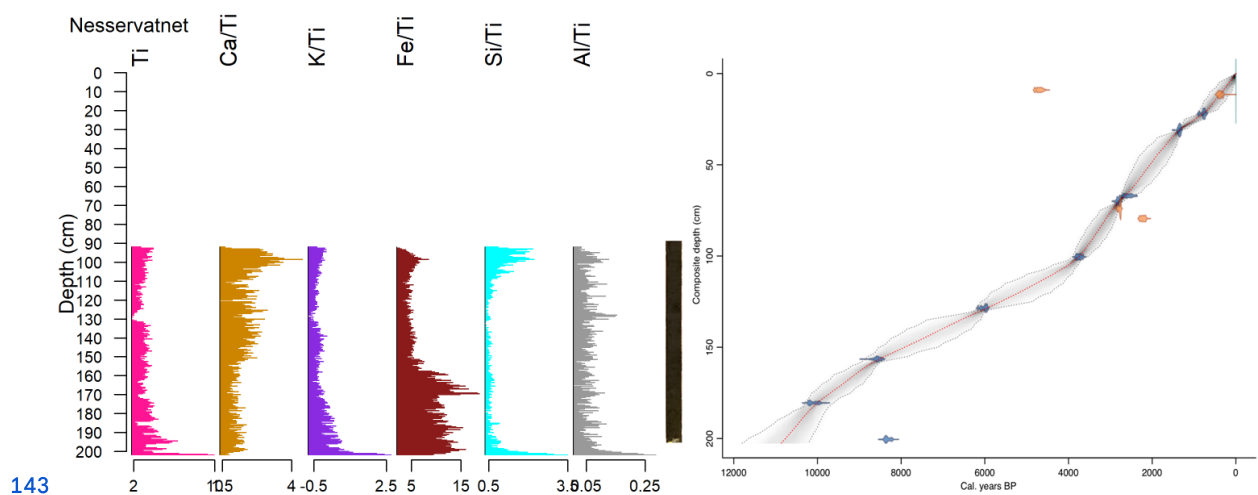
138

139



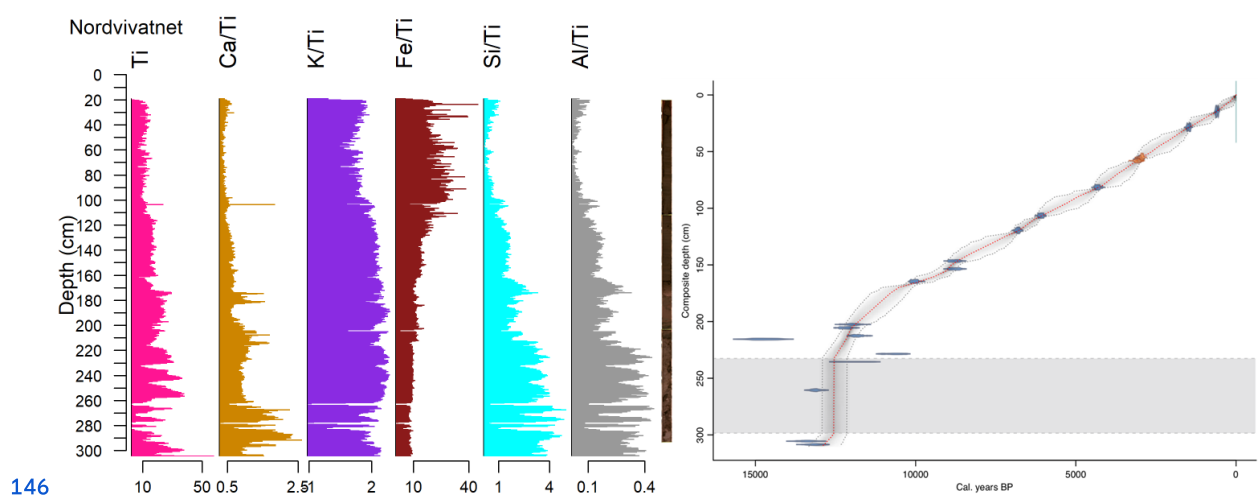
141

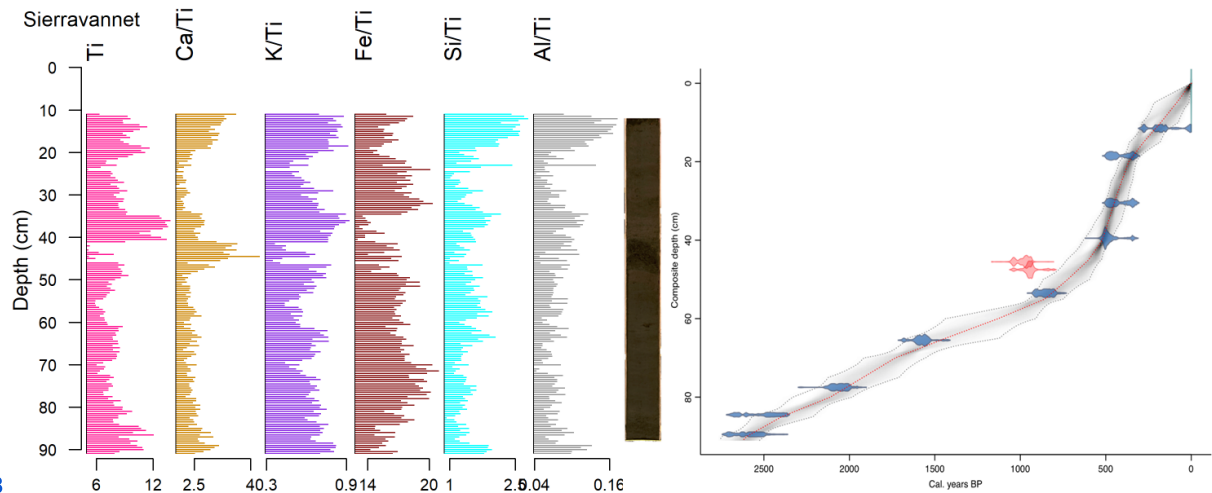
142



144

145





Supplementary Fig. 2.2: Selected elements from XRF analysis and age-depth model for Sierravannet (EG07). Raw counts of all other elements were divided by Titanium (Ti) and raw counts and Ti itself was rescaled by dividing it by 1000. Unusual high values were removed assuming core start and core breaks. This regards for 12 and 91 cm depths for Eastorjavri; 162 cm depth for Gaupstjern; and 110, 186.5 and 389 cm for Langfjordvannet. Age-deth models are taken from ²² with minor updates from ²⁸. Full XRF dataset is given as Supplementary Data 1.

3. Mitigating false positives from sampling to final dataset

To minimize the risk of false positives, we took necessary steps all along the process from coring to the final stage of bioinformatic and manual cross-checking of the data. During coring, DNA from the environment may potentially enter the sediments. To avoid this, we used a closed coring system like Nesje corer ²⁹. Before coring, we smeared/spread the pipe, piston and core head with extracted plant DNA of a Christmas cactus (*Schlumbergera truncata* cultivar) (Supplementary Fig. 3.1a), which belongs to a plant family (Cactaceae) not native to Eurasia. DNA from the same plant was also genome skimmed as part of PhyloNorway, and we have never recorded it in any of our *sedaDNA* data, suggesting that contamination during the field in general is of low risk.

During core splitting and sampling, as well as DNA extraction, all procedures are carried out sequentially from the oldest to the youngest sediment layers to prevent contamination of older layers by younger ones. Consequently, if cross-contamination were to occur during laboratory processing, it would more likely be from the older sediments to the younger ones.

176 All ancient DNA work is done in a dedicated lab for ancient DNA with over pressure.
177 All personnel working in the lab are wearing fully body suits and masks. In addition,
178 specific lab clothes that are not worn outside are used underneath the suits. All items
179 brought into the lab are washed minimum twice with a bleach solution, in addition no
180 outer packaging e.g. cardboard boxes of consumables are entering the inner lab
181 space. This to avoid potential modern contamination.

182

183 We cut the plastic of the pipe using a saw, only penetrating 0-1 mm into the 10 cm
184 diameter core. Then we cleaned the pipe with bleach before taking it into the ancient
185 DNA clean laboratory where we split the cores by pulling a string through the core
186 from the oldest to the youngest end of the core. The splitting process risks dragging
187 DNA from the core end and/or the inner core from the older to the younger sediment.
188 Any such contamination is removed by scraping off the upper sediments first once,
189 and then a second time within the first scrape to ensure that only undisturbed
190 sediments are sampled (Supplementary Fig. 3.1b). In addition, a container with DNA
191 free water is placed next to the sampling to monitor any contamination from the DNA
192 lab.

193

194 During DNA extraction, a negative control is added at the end of each batch
195 (Supplementary Fig. 3.1c). This is a completely empty sample as no water is added,
196 just controlling for cross-contamination as well as the contamination from reagents
197 as these are not produced for ancient DNA and commonly contain a small amount of
198 DNA from human and domestic species ^{30,31}.

199

200 After extraction the samples were randomly placed on a 96 well plate for easier
201 processing for PCR. During the transfer of 85ul DNA extracts from the individual
202 tubes to the plate, a negative control with molecular grade water was added to
203 account for this step. During PCR another negative control is added. In addition, we
204 included the negative controls from sampling and extraction (Supplementary Fig.
205 3.1d). By including these negative controls at all stages of the process, we can
206 monitor not only which species may occur as contaminants, but also at which stage
207 they entered the process. We did record contamination of chicken (1 negative
208 control) cow (1 negative and 1 positive control), pig (2 negative and 1 positive
209 controls) and human (36 negative controls and 11 positive controls) (Supplementary
210 Data 1). These are common contaminants found in samples analysed in different
211 ancient DNA laboratories ³¹⁻³³ and are assumed to come from the reagents and/or
212 disposables. It is possible to distinguish this type of modern contamination from
213 ancient DNA by doing shotgun analyses followed by analyses of DNA damage
214 patterns ³⁴. However, domestic animals were not the focus of this study, and the only
215 species recorded were misidentification between a domestic and wild animal is
216 possible wolf/dog. As carnivores in general occur in low number of reads (mean 24
217 reads in 60 samples, and damage pattern only found in 8 samples, ³²), we did not
218 attempt shotgun sequencing to confirm potential wolf occurrence. None of the 85

native animal taxa were detected in any of the 89 controls, supporting no physical contamination of samples.

221

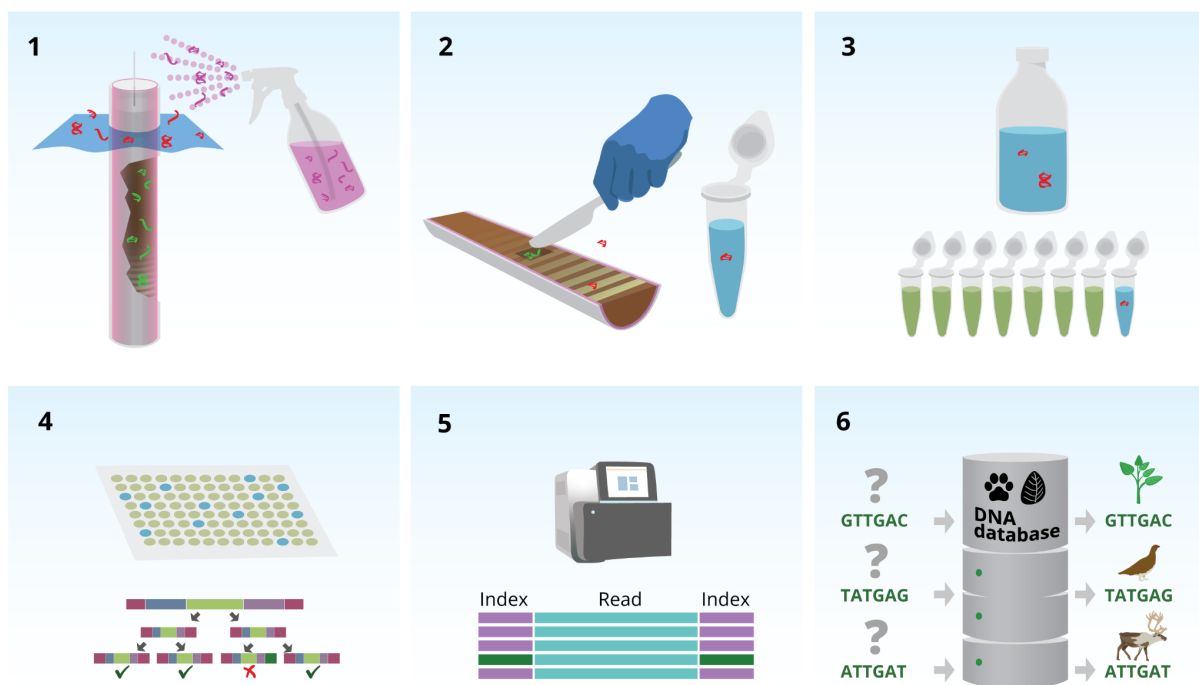
To avoid cross-over during PCR, we used dual tagging of 8-9 bases (Supplementary Fig. 3.1e). Similarly, to avoid tag jumping during sequencing³⁵, we used dual tagged library (Supplementary Fig. 3.1f). Thus, the chance that we infer the wrong date of a record due to tag jumping is low.

226

We matched the DNA sequences to the EMBL standard reference database (downloaded on 2022-10-06) which was manually curated in order to remove erroneously identified sequences. Finally, we evaluated the identified metabarcode dataset manually and removed one sequence that we were not 100% sure if it represented a wolf or contamination from dog DNA.

232

233



234

Supplementary Fig. 3. Avoiding false positives in metabarcoding at all steps from sampling to the final dataset. 1) Adding DNA tracer to equipment at coring, 2) Sampling from inner surface after double cleaning; putting out negative control next to sampling to record any background DNA in the clean laboratory, 3) Negative control of DNA-free water at the end of each batch during DNA extraction to record any cross-contamination from samples as well as background DNA from the reagents, 4) Negative controls during PCR as in step 3; dual tagging of samples to detect any cross-over during PCR, 5) Dual indexing of libraries to detect any tag jumps, 6) Matching to manually cross-checked reference library.

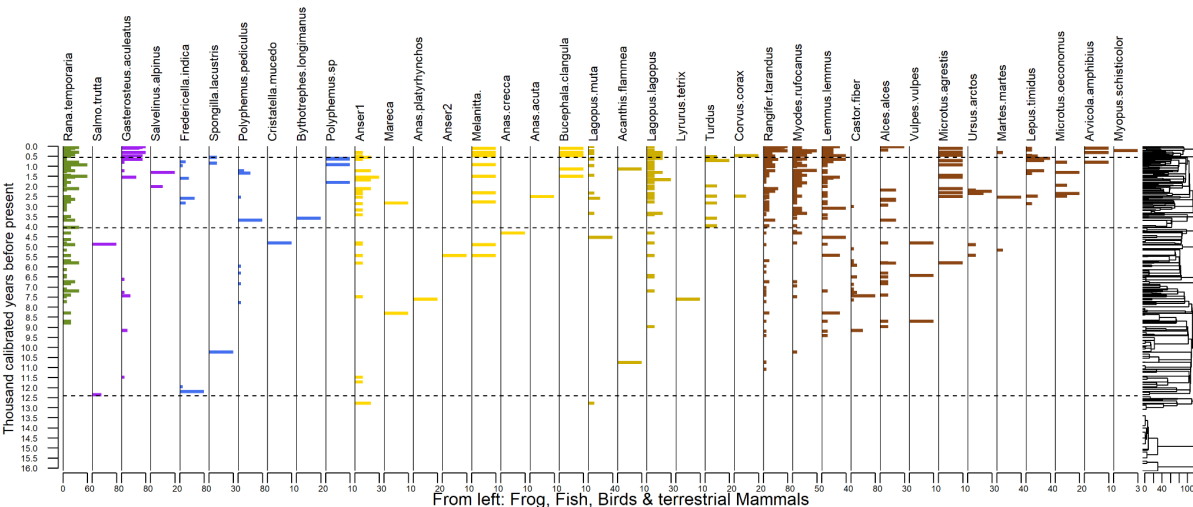
243

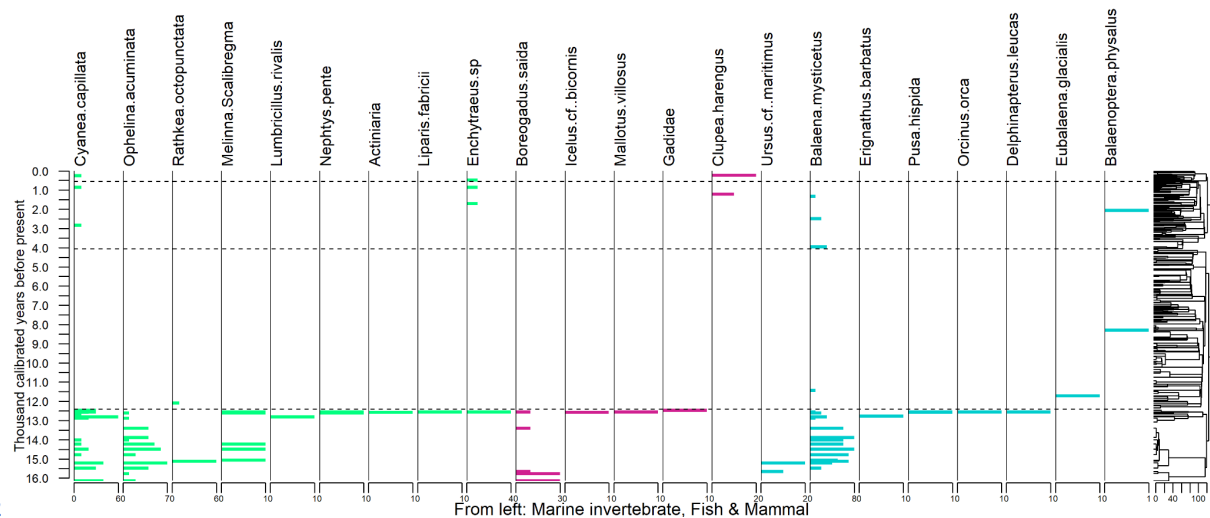
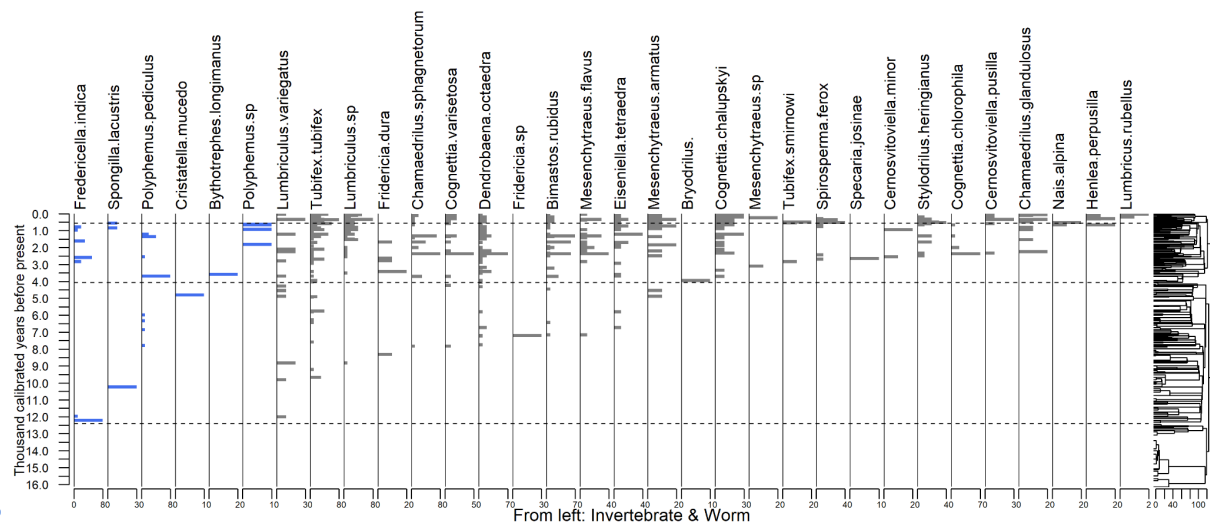
244

4. The full animal *sedaDNA* record

The full animal *sedaDNA* record includes 85 animal taxa detected in 244 of samples and is provided in Supplementary Fig. 4.1. Note that for the overview Fig.1 in the main text, the two sequences identified to *Anser* are merged, although their separate occurrences in samples suggest that they may represent different species. Note also that according to the CONISS analyses, the major division in the animal community is at 4.1 ka cal BP with further zone divisions at 12.4 and 0.55 ka cal BP.

We note that some species that are found in the region today, were not detected in *sedaDNA*. For Grey wolf, the sequence sharing with dogs challenge reliable identification (Supplementary Table S2), as dog DNA may also come from reagents (see above). Other regional species include arctic fox (*Vulpes lagopus*), Eurasian lynx (*Lynx lynx*), common shrew (*Sorex araneus*), Eurasian otter (*Lutra lutra*) and the mustelids least weasel (*Mustela nivalis*) and stoat (*Mustela erminea*). In the warmer parts of the Holocene, roe deer (*Capreolus capreolus*) would also be possible. As these taxa are all represented in the DNA reference library, the lack of detection is not due to the lack of a reference. Bones of Eurasian otter are recorded in the region (Supplementary Table S4) and the lack of *sedaDNA* could for most sites be explained by the distance to the sea. Mountain hare and European water vole were detected earlier in bones than in *sedaDNA*, and the earliest records of reindeer and hare were surprisingly late compared to its early arrival in western Norway³⁶. The other species are also lacking from the regional bone records (Supplementary Table S4). For example, lynx are also rare in archaeological records, and their past distribution is therefore largely unknown³⁷. Overall, we detected mammals in 58% of the sample with on average 0.94±1.15 species per sample (63% of samples last 6 k with 1.25±1.26 species). This is similar to what has been found overall for a circumpolar study of 495 samples (1.86 taxa, 63% of samples), and higher than found in lake sediments (38% of samples with 0.69 taxa,³² suggesting that our methods worked well. Thus, the low carnivore detection may be related to low biomass abundance, leaving only scattered *sedaDNA* that is challenging to detect.

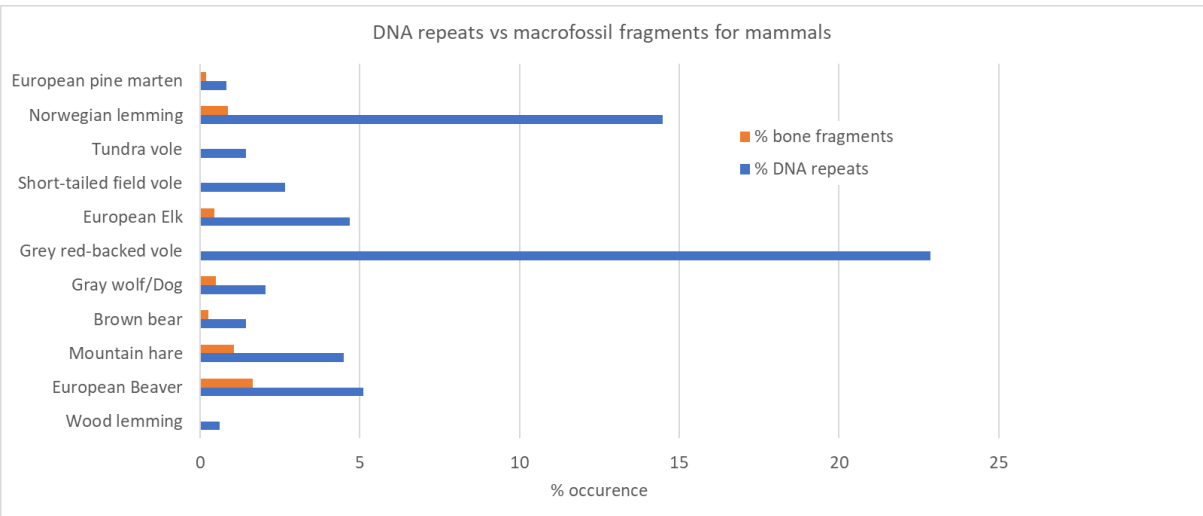
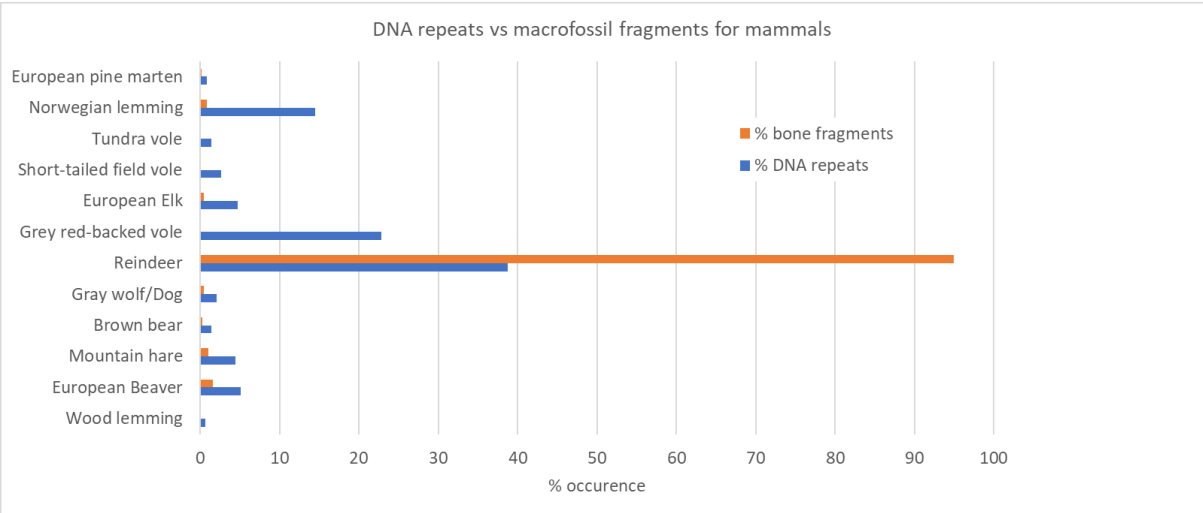




Supplementary Fig. 4.1 Identification of animals in ten sediment cores from N Fennoscandia. The quantification is given as numbers of PCR repeats (1-8); note the difference in scale between taxa. **a** Terrestrial and freshwater vertebrates (green=amphibian, purple=fish, yellow=bird, brown=mammals). **b**, Freshwater invertebrates (blue) including worms (grey, mostly assumed to be freshwater/terrestrial). **c**, Marine organisms (green=invertebrate, purple=fish, and blue=mammal).

5. Comparison of bone and sedaDNA records

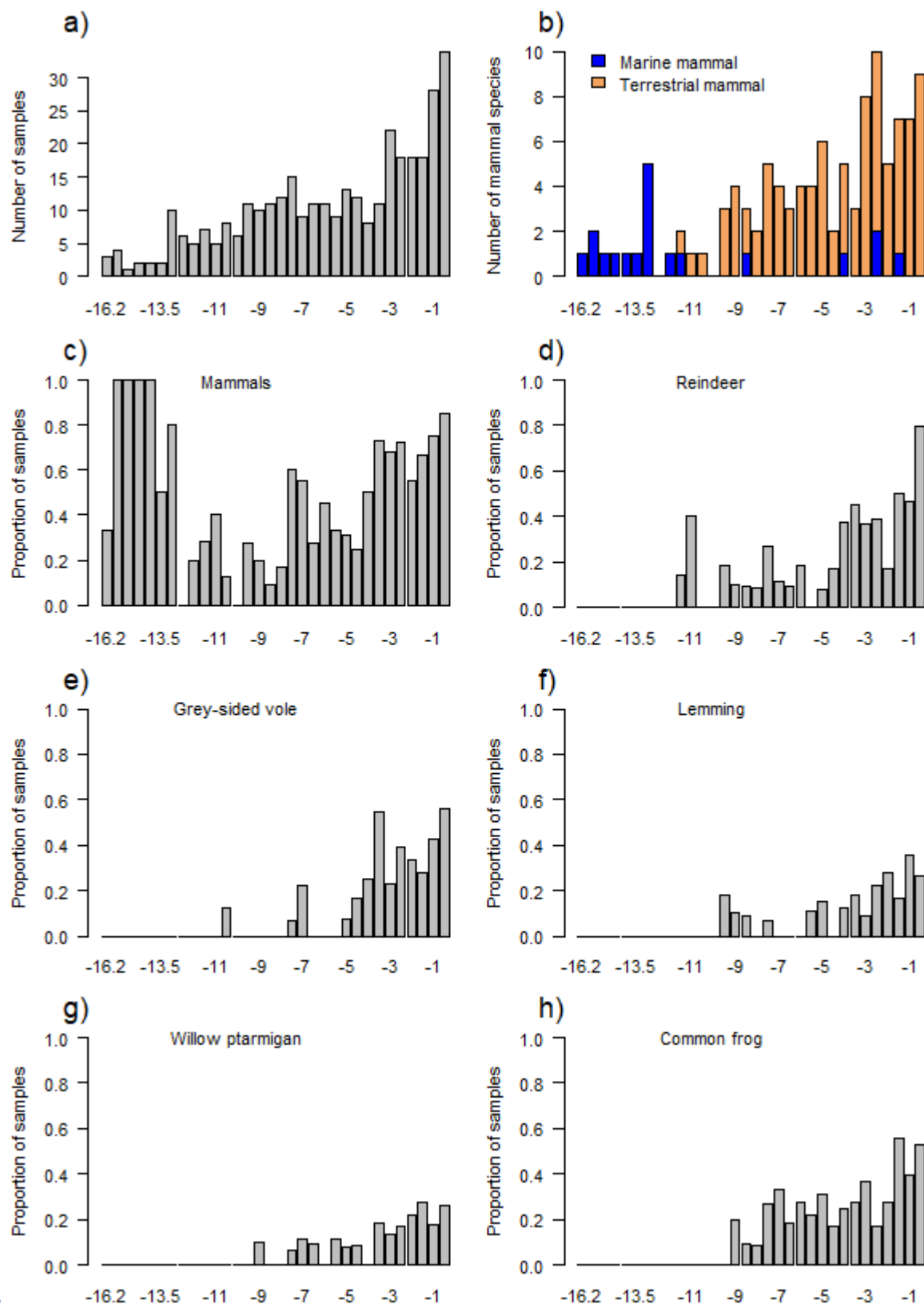
Of the 12 mammal taxa identified, all were recorded in sedaDNA and eight were recorded as bones. Note that quantification is uncertain based on both methods, but the bone records were clearly strongly biased towards reindeer (Supplementary Fig. 5.1), which, based on the bone record, is assumed to have been a major part of the human diet in the region³⁸.



Supplementary Fig. 5.1 Relative abundance of taxa detected in sedaDNA (% of repeats) and bone fragments (% of fragments). **a**, All mammals. **b**, Excluding reindeer.

6. Animal frequency over time

Both numbers of samples and numbers of animals detected increased towards the present (Supplementary Fig. 6.1a-b). However, the proportion of samples that contained the total record of mammals, or the individual species (reindeer, Norwegian lemming, grey-sided vole, willow ptarmigan and frog) also increased towards the present (Supplementary Fig. 6.1c-h), showing that the increase is not only due to sampling effort. Moreover, as the sedaDNA quality was stable over this time period²², we assume that the increase in detection represents a true increase in occurrences, as also seen by the increase in sites where reindeer was detected (Supplementary Fig. 6.2).

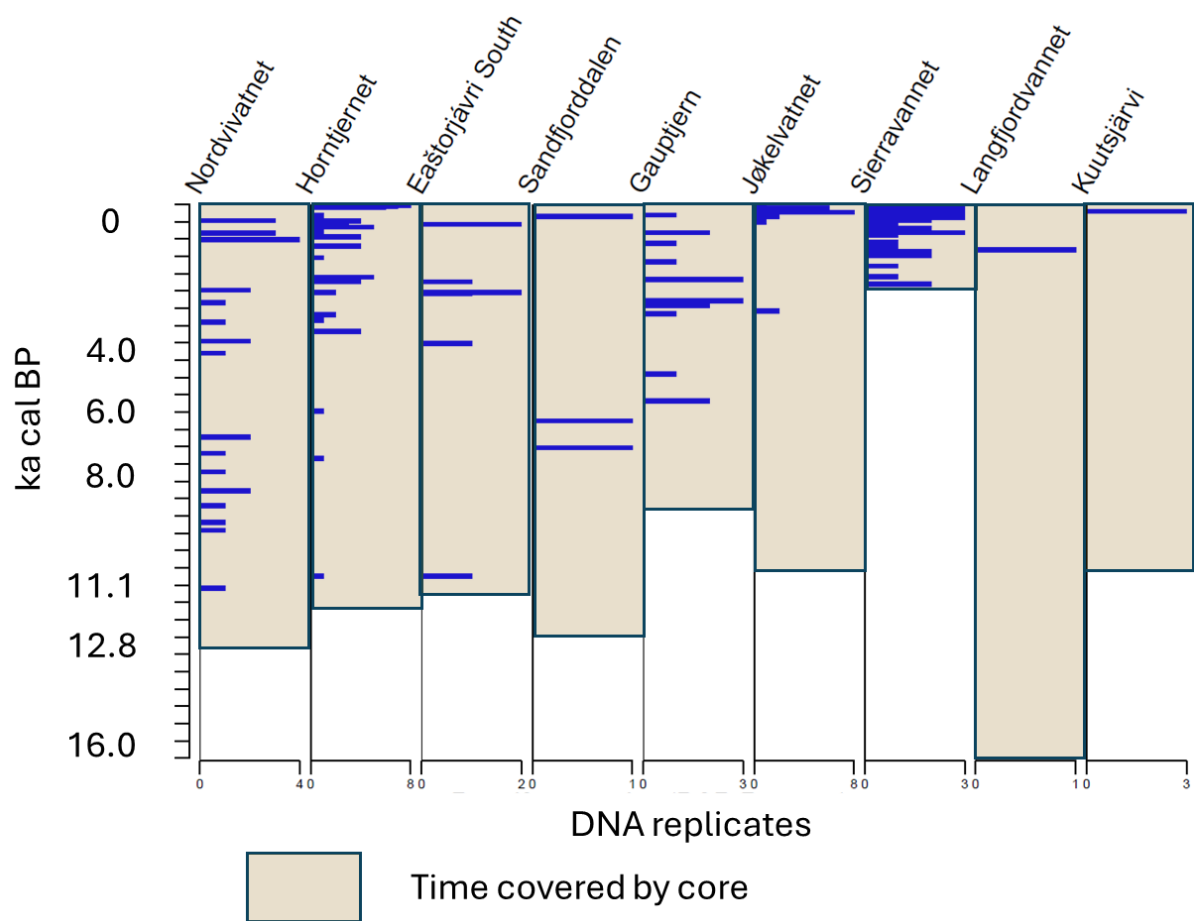


316

317

318 *Supplementary Fig. 6.1 Density of sampling and detection of animals in sedaDNA analysis*
 319 *per 500 year time interval. a, Number of samples. b, Number of mammals species detected.*
 320 *Proportion of samples with c, total mammals, d, reindeer, e, grey-sided voles, f, lemmings,*
 321 *g, willow ptarmigan, and h, frog.*

322



323

324 *Supplementary Fig. 6.2. Reindeer appearance at the nine sites it was recorded (there were no*
325 *records of reindeer from Nesservatnet on the island Årøya). Lake Jøkelvatnet is on a*
326 *peninsula where dispersal is limited by a glacier and steep hillsides, Langfjordvatnet is on the*
327 *northern island of Arnøya, and lake Kuutsjävri is presently dominated by spruce and pine*
328 *forest.*

329

330

331 7. Plant-herbivore co-occurrence networks

332 We plotted co-occurrence networks for significant co-occurrences across three methods only
333 (Fig. 2), and for 75% co-occurrences (Supplementary Fig. 7.1). We show all 1474
334 co-occurrence in Supplementary Table 6 along with the plant traits and significant levels
335 according to our different tests. We are not aware of any similar data even for the present
336 that shows plants-herbivores co-occurrence, but the patterns fit with the habitat description
337 of individual species as for example described in the Norwegian encyclopaedia Store Norske
338 Leksikon (<https://snl.no/>).
339

340 7.1 Reindeer co-occurrence

341 Reindeer showed both significant and >75% co-occurrence with *Empetrum nigrum*,
342 *Rubus/Fragaria* (presumably *Rubus chamaemorus*), *Juniperus communis* and *Pinus*
343 *sylvestris*. These plant taxa appeared before reindeer, although *Pinus sylvestris* did not
344 appear regularly until 10.3 ka cal BP and *V. myrtillus* appeared 10.5 ka cal BP
345 (Supplementary Table S6, ²⁸). Most of these taxa are known as common food plants to
346 reindeer ^{39,40}. Reindeer further showed >75% co-occurrence the common food plants
347 *Vaccinium myrtillus*, *V. uliginosum*, *V. vitis-idaea*, *Betula*, and *Salicaceae*. The plants
348 co-occurring with reindeer have generally wide habitats, but there is a high proportion of
349 graminoids and woody species mainly growing in sub-alpine or forest.
350

351 7.2 Grey-sided vole co-occurrence

352 There were 21 plant taxa with >75% co-occurrence with grey-sided vole, 12 of them
353 significant. This included tall forbs such as *Angelica archangelica*, *Trollius europaeus*, and
354 *Filipendula ulmaria* as well as several ferns and trees and small forbs. All except three of
355 these were present prior to first detection of grey-sided vole. There were 37 additional plant
356 taxa showing significant co-occurrence with grey-sided vole, 22 of them arriving after the first
357 appearance of grey-sided vole. Important food plants as *Empetrum nigrum*, *Vaccinium*
358 *uliginosum*, *V. myrtillus*, and *V. vitis-idaea* were identified as significant in two of the three
359 approaches whereas *Betula* and *Salicaceae* were co-occurring in almost all samples but not
360 significantly. The 75% co-occurring plant taxa represent all growth forms and a wide range of
361 habitats from arctic-alpine like *Oxyria digyna* and *Harrimanella hypnoides* to boreal like
362 *Pinus sylvestris*, *Linnaea borealis* and *Matteuccia struthiopteris*.
363

364 7.3 Norwegian lemming co-occurrence

365 There were 14 plant taxa that showed 75% or more co-occurrence with Norwegian lemming,
366 but only one of them significantly according to all three test (*Gymnocarpium*). Other taxa with
367 75% co-occurrence included *Vaccinium uliginosum*, *Empetrum nigrum*, *Betula*,
368 *Gymnocarpium*, *Carduinae*, *Juniperus communis*, *Rubus/Fragaria* (presumably *Rubus*
369 *chamaemorus*), *Solidago virgaurea* and *Saussurea/Lactuca*. The former four are also found
370 in Norwegian lemming diet ^{41,42}. There were a further five plant taxa with >75%
371 co-occurrence of which *Salicaceae*, *Vaccinium myrtillus*, *V. vitis-idaea* and *Bistorta vivipara*
372 are known food items ^{41,42} and *Sorbus aucuparia* a likely food item. A further 26 plant taxa

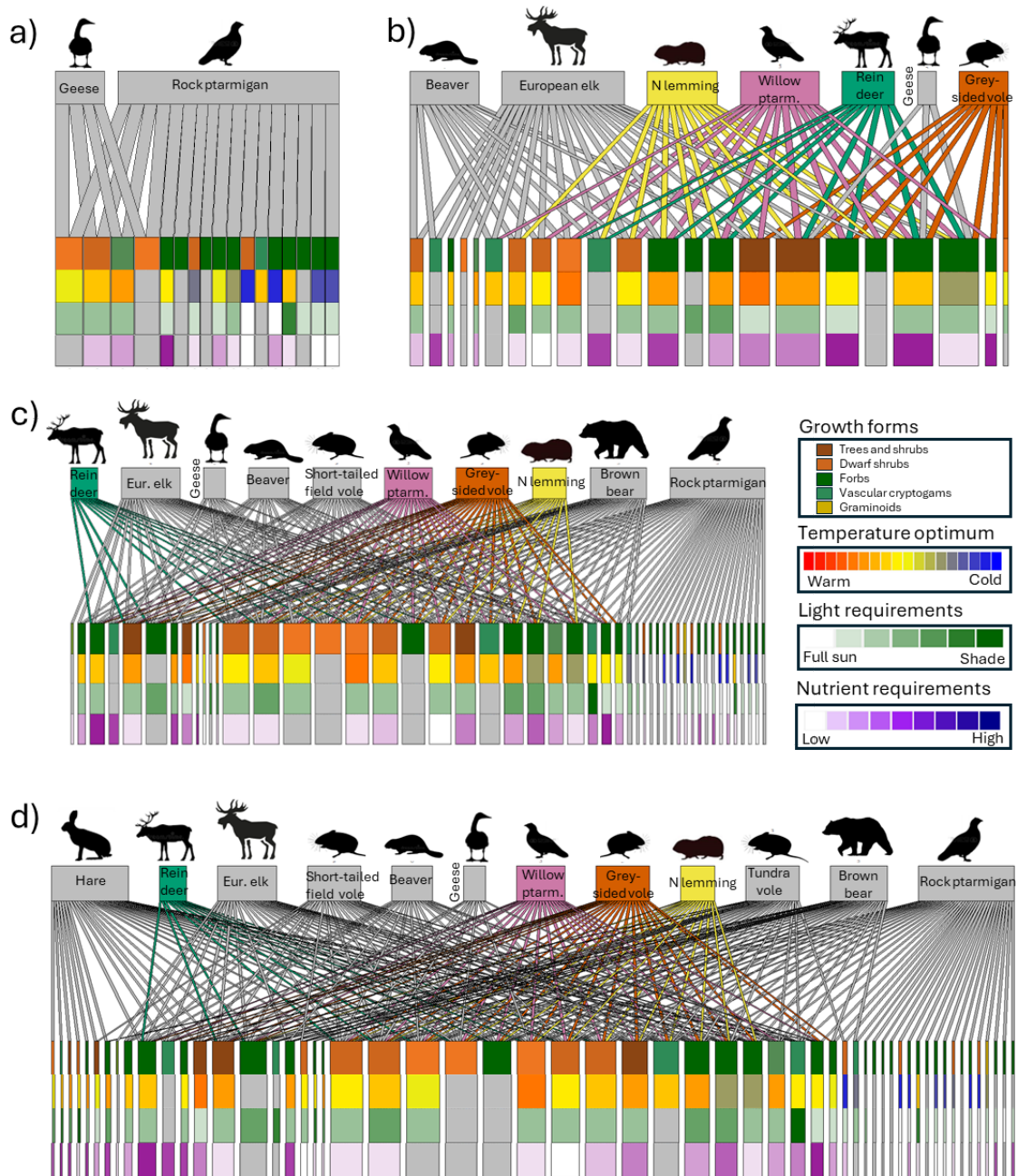
373 showed significant co-occurrence according to all three test. These include food plants like
374 *Rumex* and several graminoids, and overall a large species typical for low to high alpine
375 sites like *Silene acaulis*, *Diapensia lapponica* and *Orojuncus trifidus*.

376

377 7.4 Willow ptarmigan

378 There were 19 plant taxa that showed 75% or more co-occurrence with willow ptarmigan,
379 This included many species with berries: *Sorbus aucuparia*, *Juniperus communis*,
380 *Vaccinium vitis-idaea*, *V. myrtillus*, *V. uliginosum*, *Empetrum nigrum* and *Rubus/Fragaria*
381 (presumably *Rubus chamaemorus*). Further, tall (*Angelica archangelica*, *Filipendula*
382 *ulmaria*) and small (*Bistorta vivipara*, *Solidago cf virgaurea*, *Saussurea alpina*) and a small
383 fern (*Gymnocarpium*), as well as the always present *Betula* and *Salicaceae*. The *Vaccinium*
384 species and *Empetrum* are also found in eDNA analyses of diet in the region, whereas
385 *Bistorta* is an important food item in general for ptarmigan^{43,44}. A further 17 taxa showed
386 significant co-occurrence according to all three tests (Supplementary Table S6). T. Of the 22
387 plant taxa that showed significant co-occurrence, 9 arrived after willow ptarmigan.

388



391

392

393 **Supplementary Fig. 7.1. Herbivore network for 75% co-occurrence. a, Late Glacial**
 394 **(12.8-11.7 ka BP). b, Early Holocene (11.1-8.3 ka BP). c, Middle Holocene (8.2-4.259 d,**
 395 **Late Holocene (≥ 4.25 ka cal BP). Animal symbols as in Fig. 1. The omnivore brown bear is**
 396 **here plotted as a herbivore.**

397

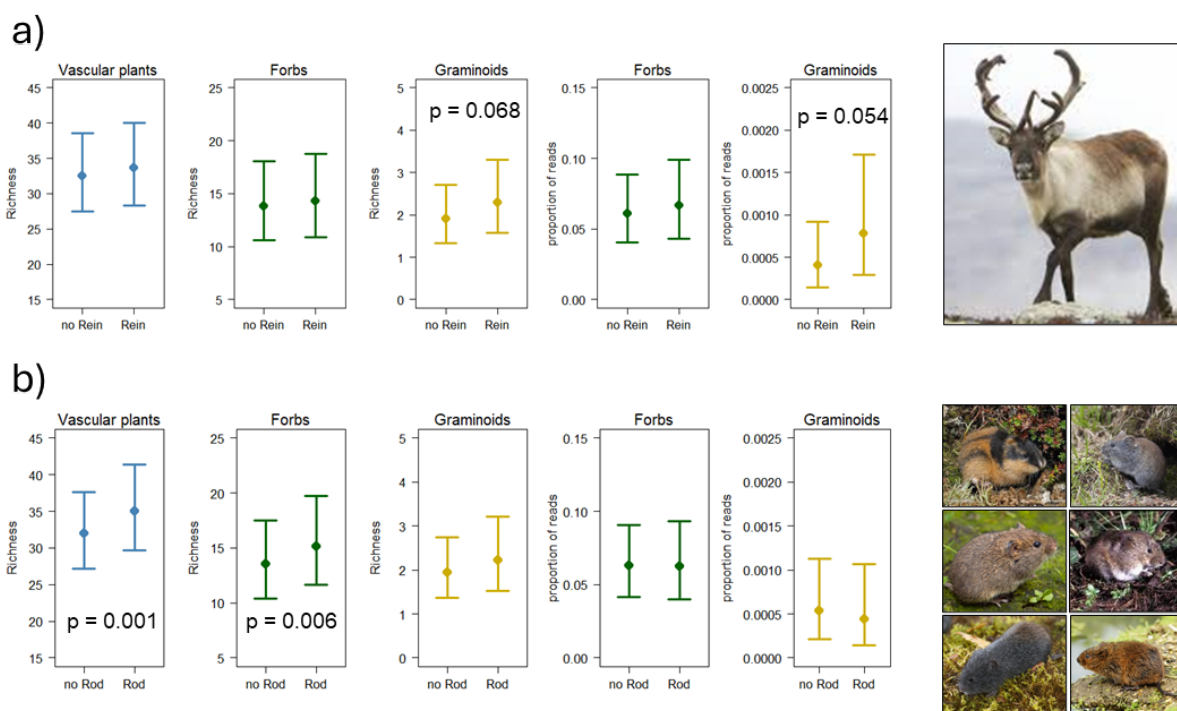
398

399

400

7.5 Strength of plant-herbivore associations

To further explore whether there were consistent differences in the vegetation in samples where herbivores were present or not detected, we tested for differences in plant species richness (number of terrestrial vascular plants detected) between samples with and without reindeer or small rodents, as well as differences in the relative abundance (proportion of reads) of forbs and graminoids in the vegetation. Both forbs and graminoids are preferred food items of herbivores and plant groups that have been shown to benefit from herbivory. Reindeer were marginally associated with plant communities with higher richness and occurred with slightly higher abundance of graminoids (richness: 1.199; Standard error SE = 0.100; 95% confidence interval CI = -0.015 - 0.379; $z = 1.807$; $p = 0.071$. Proportion of reads: 0.027, SE = 0.013, 95% CI = 0.001 - 0.053; $z = 2.070$, $p = 0.039$; Supplementary Fig. 6.5a), whereas the six small rodents taken together were associated with plant communities with higher vascular plant richness, especially forb richness (vascular plant richness: 0.089; SE = 0.028; 95% CI = 0.033 - 0.144; $z = 3.14$, $p = 0.002$. Forb richness: 0.115; SE = 0.042; 95% CI = 0.032 - 0.197; $z = 2.732$; $p = 0.006$) Supplementary Fig. 6.5b). The difference in mean species richness values was low, suggesting that the interaction between plants and herbivores was not a very important driver of plant richness patterns (Fig. 1a). This finding is in strong contrast to the Alps, where a strong increase in species richness, especially the forb component of plant communities, was reported in response to the introduction of domesticated mammals⁴⁵. Northern Fennoscandia is the region with the greatest number of studies on Arctic contemporary herbivores, and the majority of those show low impact of herbivores on plant community compositions, likely because the winter conditions keep herbivores at low densities⁴⁶. Thus, our findings of past plant-herbivore associations confirm the general findings based on contemporary studies.



Supplementary Fig. 7.5 Relationships between herbivore presence and plant abundance and richness for selected plant growth forms (forbs and graminoids). a, Impact of reindeer

429 (*Rein*; *Rangifer tarandus*), and **b**, impact of all small rodents (*Rod*) found in the region. From
 430 the top left: Norwegian lemming (*Lemmus lemmus*), grey-sided vole (*Myodes rufocanus*),
 431 short-tailed field vole (*Microtus agrestis*), tundra vole (*Microtus oeconomus*), wood lemming
 432 (*Myopus schisticolor*) and European water vole (*Arvicola amphibius*).

433

434 8. Niche change for individual species

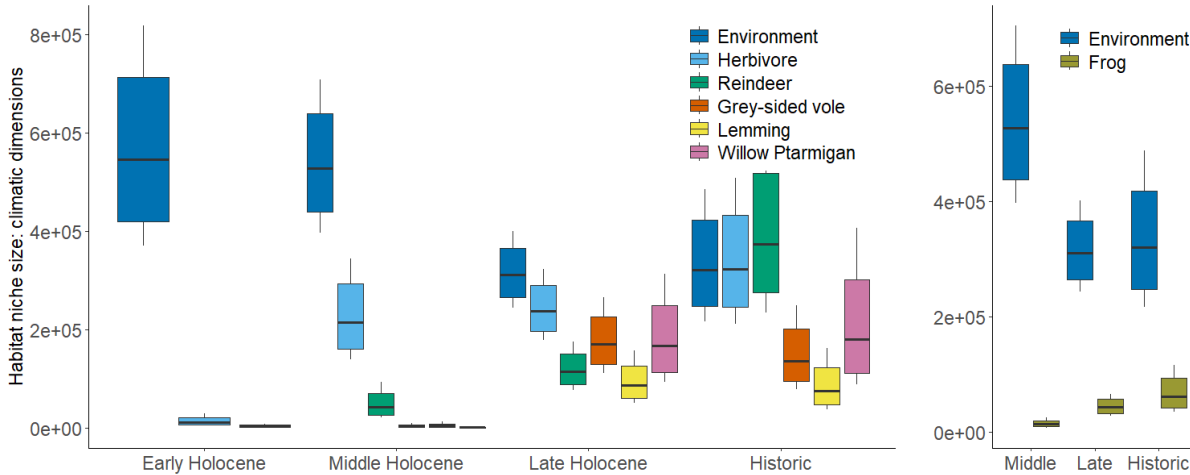
435

436 8.1 Changes in niche sizes

437 Both niche size (Supplementary Fig. 8.1) and niche dimensions (Supplementary Figs.
 438 8.2-8.6) changed over time. Below, we provide some comments on the individual species,
 439 whereas the overall pattern is displayed and described in the main text.

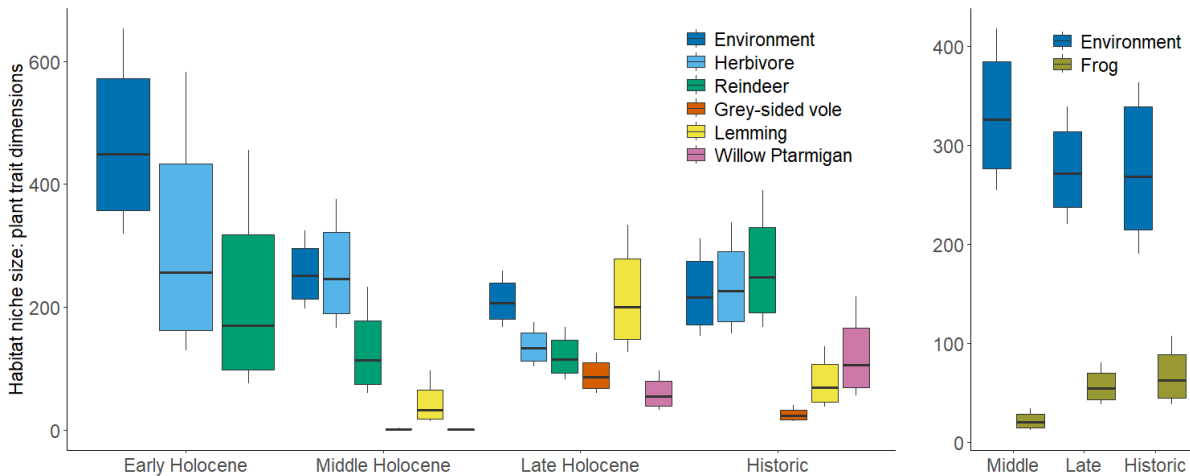
440

441 a)



442

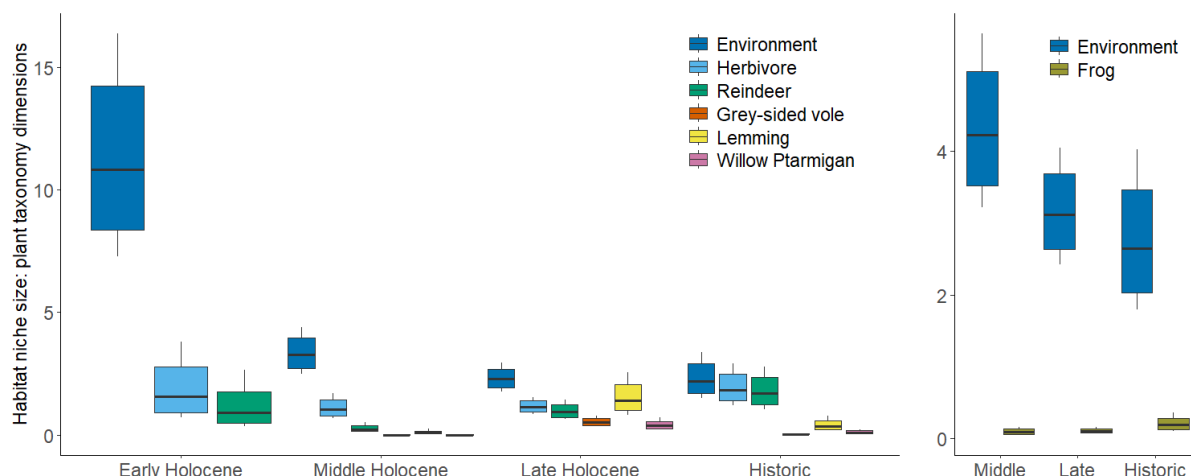
443 b)



444

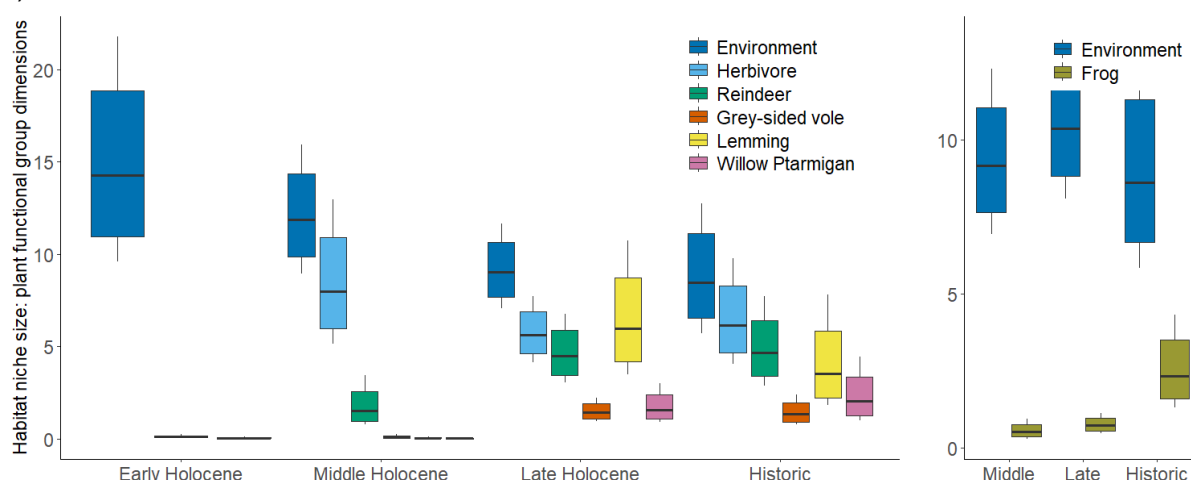
445

446 c)



447

448 d)



449

450

451 *Supplementary Fig. 8.1. Niche size for different niche dimensions and time periods for each*
 452 *of five species. a, Climatic niche, b, abiotic plant traits, c, plant taxonomic diversity, e, plant*
 453 *functional groups.*

454 8.2 Reindeer niche

455 The single niche dimension that predicted reindeer occurrence best was annual
 456 precipitation, followed by taxonomic diversity NMDS dimension1, July temperature and
 457 functional diversity NMDS dimension 4 (Fig.3a). The set of niche dimensions that explained
 458 the occurrence of reindeer best was climate (44%) followed by plant taxonomic diversity
 459 (26%), whereas plant abiotic traits and plant functional groups explained 14-15%. Reindeer
 460 mainly occurred in samples with annual precipitation below 550 mm and July temperature of
 461 10-15°C.

462

463 There were large changes in niche dimensions, especially the climate niche, with less than
 464 20% probability of overlap among 8 of the 12 comparisons among time periods
 465 (Supplementary Fig. 8.2a). There was less shift in abiotic traits based on plants, with about
 466 90% probability of the niche from Late Holocene being found within the present day niche

467 (Supplementary Fig. 8.2b). Similarly, the largest shifts in plant taxonomic diversity and plant
468 functional traits were from Early to Middle and Middle to Late Holocene, with a smaller shift
469 from Late Holocene to the Historical time period (Supplementary Fig. 8.2c, d). When all 15
470 niche dimensions were analysed together, the largest change in niche was between the two
471 most recent periods compared to the two oldest periods (Supplementary Fig. 8.2e).

472

473 The size of the climatic and abiotic niche are significantly increasing in the last 650 years,
474 probably because humans brought the reindeer to new sites where they do not occur
475 naturally. This is further confirmed by their occupation of almost all environmental space for
476 climate, abiotic traits and plant taxonomic diversity, whereas their niche sizes in earlier
477 periods only occupy a fraction of the environmental spaces (Fig. 3).

478

479 Reindeer have large ranges and show annual migration between different habitats. Their diet
480 is broad, with a high number of vascular plants, bryophytes and lichens identified ^{40,47}, and
481 their most common food items are widespread in the region. In contrast, their heat tolerance
482 is low ⁴⁸ and the accessibility of food may be limited by snow and ice during winter ⁴⁹. This
483 might explain why climate niche dimensions seem more important than vegetation for
484 reindeer.

485

486

487

488

489

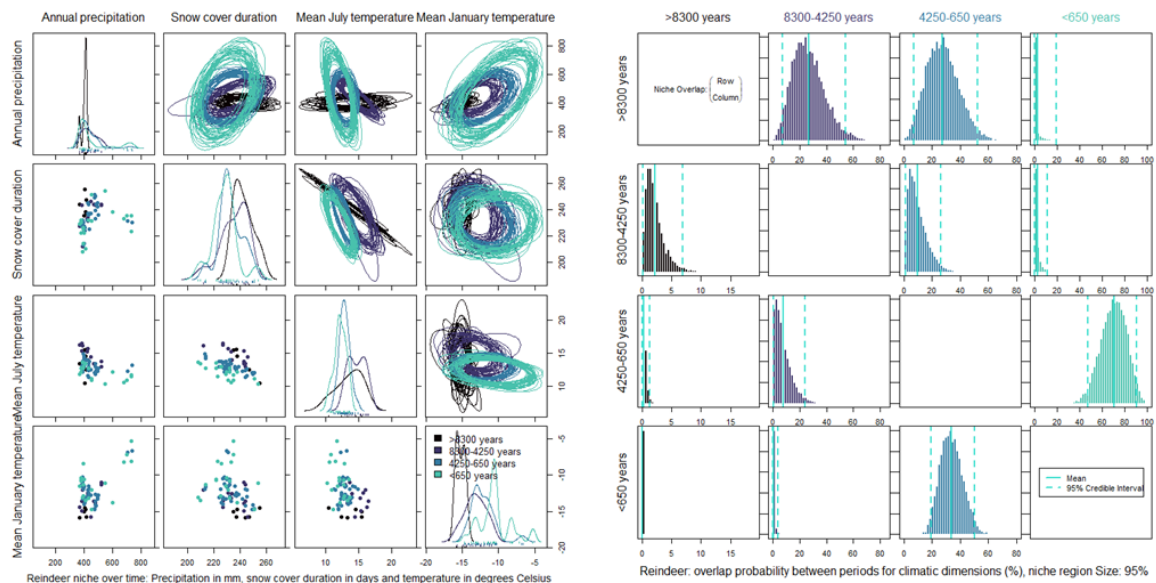
490

491

492

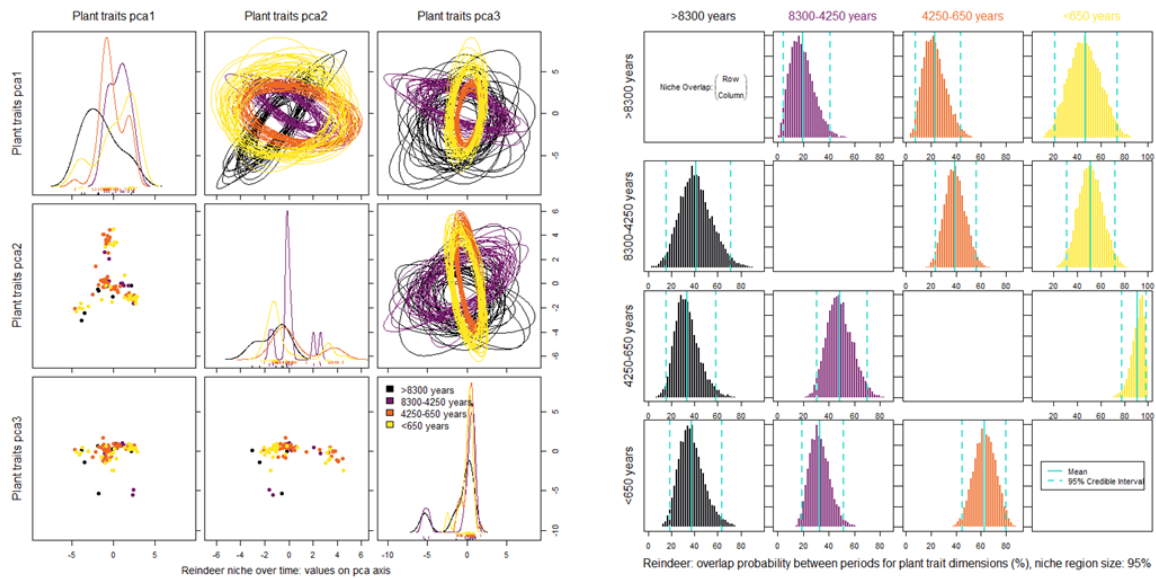
493

494a)



495

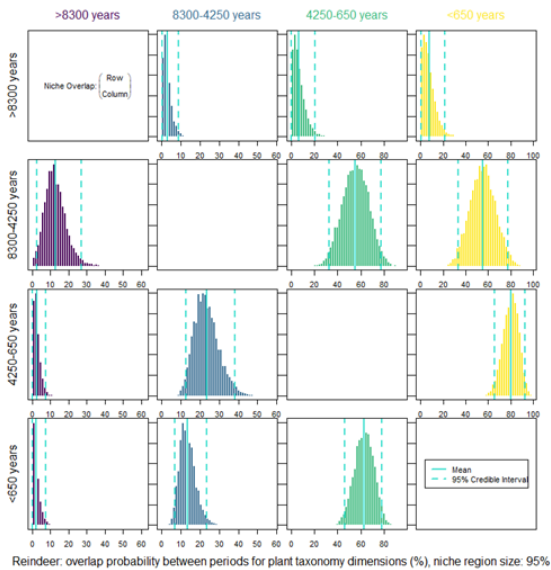
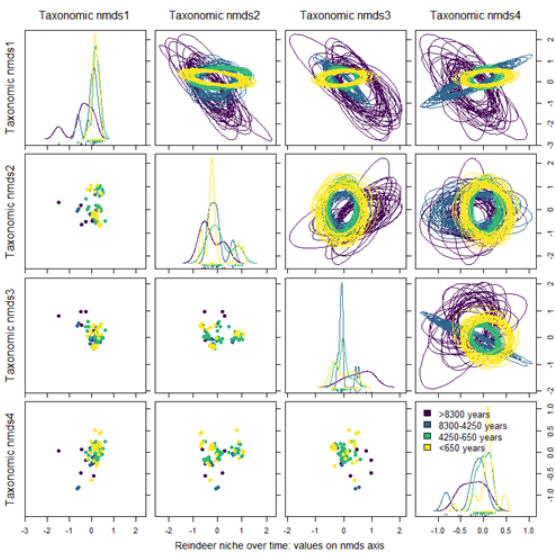
496b)



497

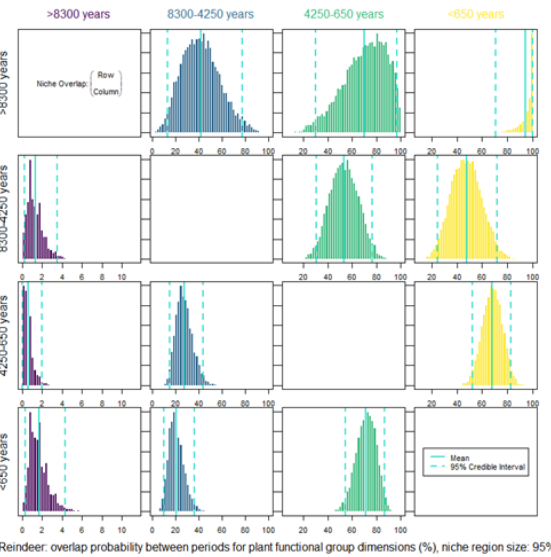
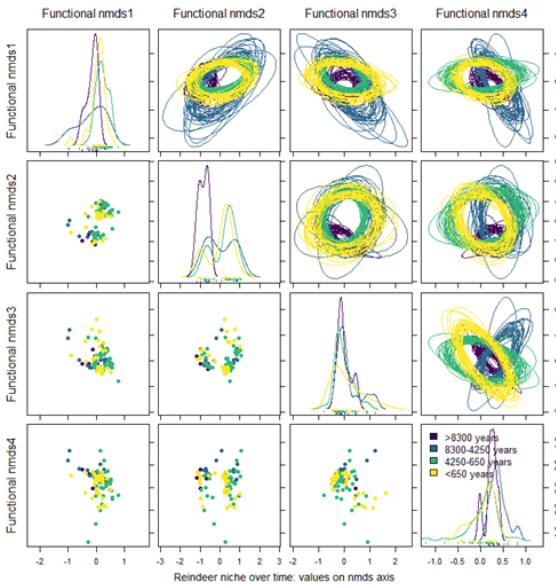
498

499c)



500

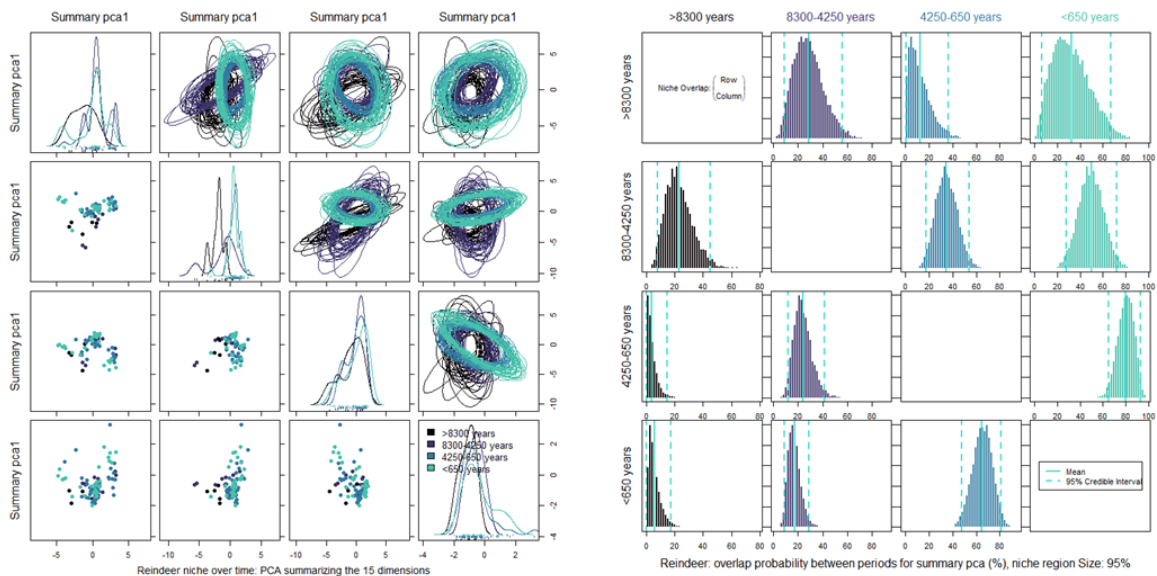
501d)



502

503

504 e)



505

506

507 *Supplementary Fig. 8.2. Niche stability over time for **reindeer** (*Rangifer tarandus*) based on*
 508 *a, climate niche dimension, b, climate and other abiotic conditions inferred from the traits of*
 509 *plants present, c, vegetation attributes with emphasis on taxonomic richness, d, vegetation*
 510 *attributes using with emphasis on abundance of plant functional composition, e, all 15 niche*
 511 *dimensions summarised using 3 PCA axis representing 70% of the variation. The bar plots*
 512 *represent the posterior probability that an individual from the time period indicated by the row*
 513 *will be found within the niche of the time period indicated by the column header.*

514

515 8.3 Grey-sided vole niche

516 Grey-sided vole occurrence was best predicted by taxonomic diversity NMDS dimension 1
 517 followed by abiotic traits dimension2 and taxonomic diversity NMDS dimension 2. The niche
 518 that best explained its occurrence was taxonomic dimensions (48%), followed by plant
 519 abiotic traits (24%), plant functional groups (15%) and climate (13%) (Fig. 3).

520

521 There were large changes in all niche dimensions over time. The probability of finding the
 522 niche of Late Holocene or Historical period in the Middle Holocene was close to zero
 523 whereas there was less change in climate niche between Late Holocene and the Historical
 524 period (Supplementary Fig. 8.3a). A very similar pattern was seen for abiotic plant traits,
 525 plant taxonomic diversity and plant functional groups (Supplementary Fig. 8.3b-d) and all 15
 526 niche dimensions combined (Supplementary Fig. 8.3e). The probability to find the Historical
 527 niche dimensions in Late Holocene was medium for climate (~50%), and high for the other
 528 sets of niches and the combination of all 15 niche dimensions (60-85%).

529

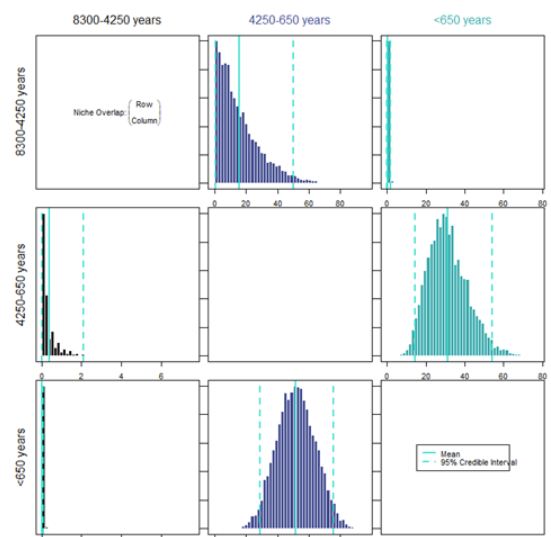
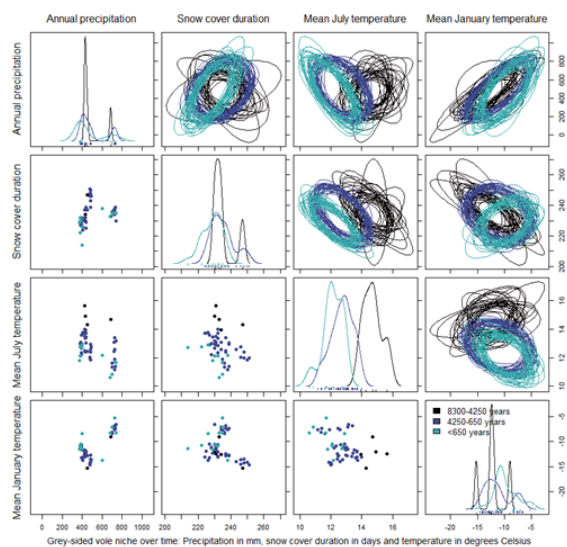
530 The size of the plant abiotic traits and plant taxonomic diversity niche was largest in Late
 531 Holocene and decreased the last 650 years (Supplementary Fig. 8.1). This is surprising,
 532 given that the proportion of samples with grey-sided voles increased towards the present,
 533 but we do note that the frequency was also high around 4000 years ago (Supplementary Fig.
 534 6.1e). The climatic niche showed minor change over the same period, suggesting that the

535 reduction is related to biotic interactions. A reduction could potentially be caused by human
536 mediated range expansion and population increase of reindeer, as these species show
537 increased overlap over time (see below). This contradicts contemporary co-occurrence
538 studies based on faeces counts that suggest that reindeer may facilitate small rodents ⁵⁰,
539 although rodents were not distinguished to species, and therefore it is uncertain if this
540 applies to grey-sided vole. The increase in niche overlap may not exclude other causes for
541 niche size reduction in grey-sided vole, for example increased predation or competition from
542 other rodents.

543

544

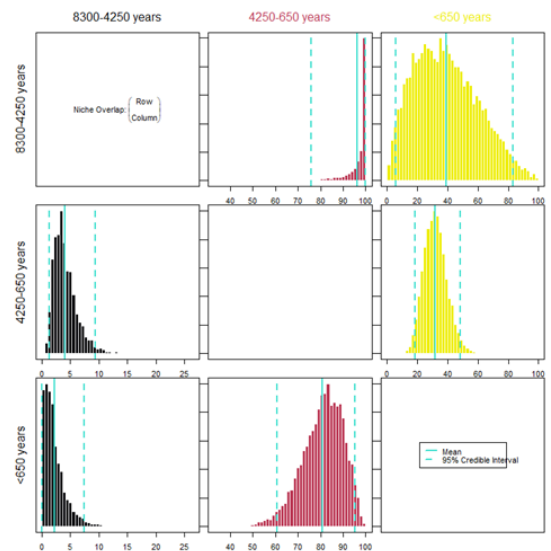
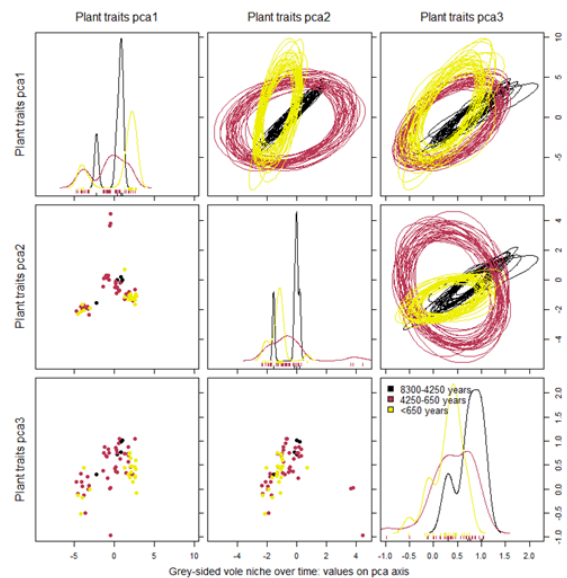
545a)



Grey-sided vole: overlap probability between periods for climatic dimensions (%), niche region size: 95%

546

547b)

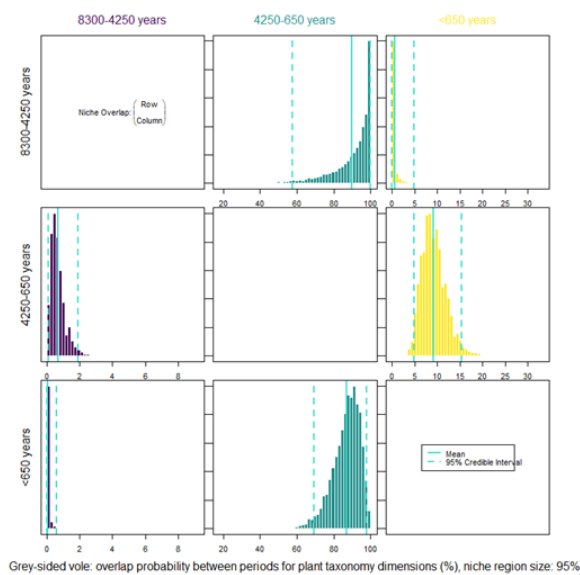
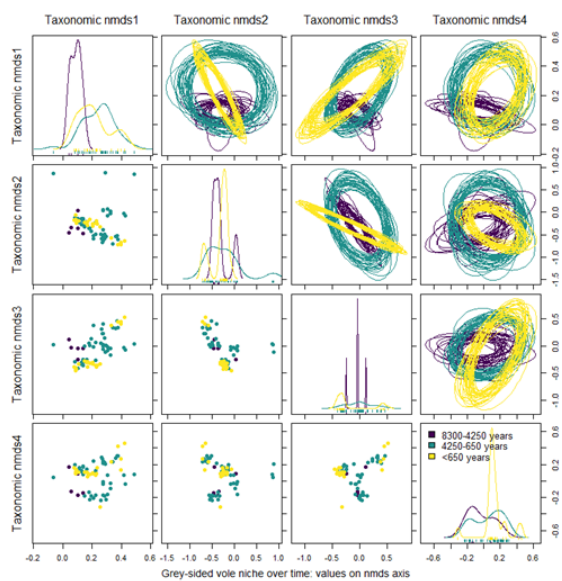


Grey-sided vole: overlap probability between periods for plant trait dimensions (%), niche region size: 95%

548

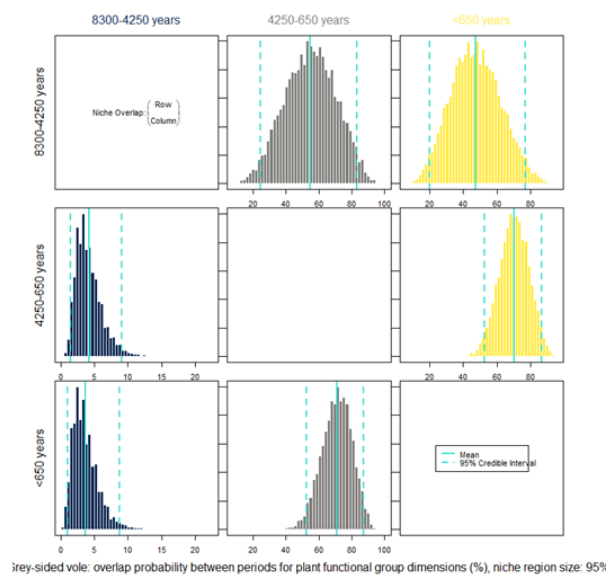
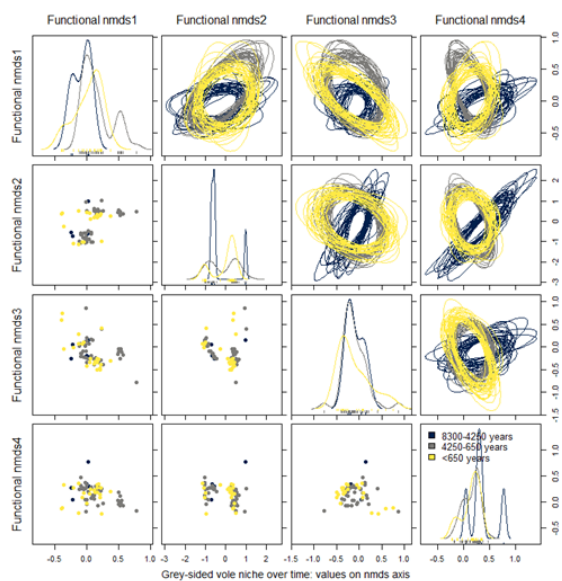
549

550c)



551

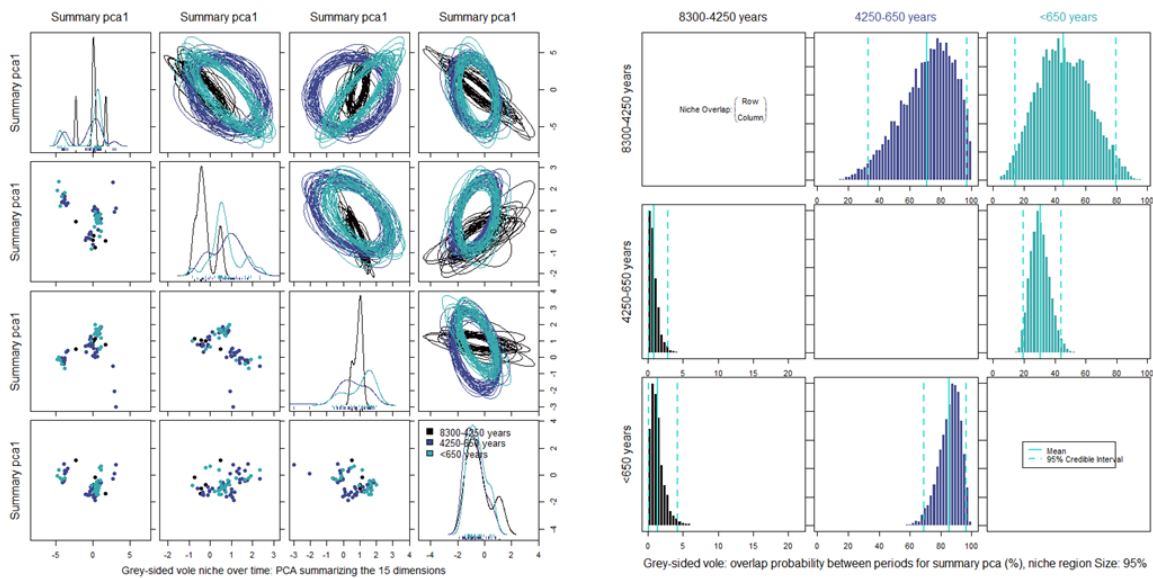
552d)



553

554

555e)



556

557

558 **Supplementary Fig. 8.3. Niche stability over time for *grey-sided vole* (*Myodes rufocanus*)**
 559 **based on a, climate niche dimension, b, climate and other abiotic conditions inferred from**
 560 **the traits of plants present, c, vegetation attributes with emphasis on taxonomic richness, d,**
 561 **vegetation attributes using with emphasis on abundance of plant functional composition, e,**
 562 **all 15 niche dimensions summarised using 3 PCA axis representing 70% of the variation.**
 563 **The bar plots represent the posterior probability that an individual from the time period**
 564 **indicated by the row will be found within the niche of the time period indicated by the column**
 565 **header.**

566

567 8.4 Norwegian lemming niche

568 The occurrence of Norwegian lemming was best predicted by plant taxonomic richness
 569 NMDS dimension 2, followed by annual precipitation, plant taxonomic diversity NMDS
 570 dimension 1 and snow cover. Overall, the occurrence of Norwegian lemming was marginally
 571 better predicted by plant taxonomic diversity (37%) than climatic niche dimensions (33%),
 572 whereas plant functional groups and abiotic traits predicted 19% and 11%, respectively (Fig.
 573 3).

574

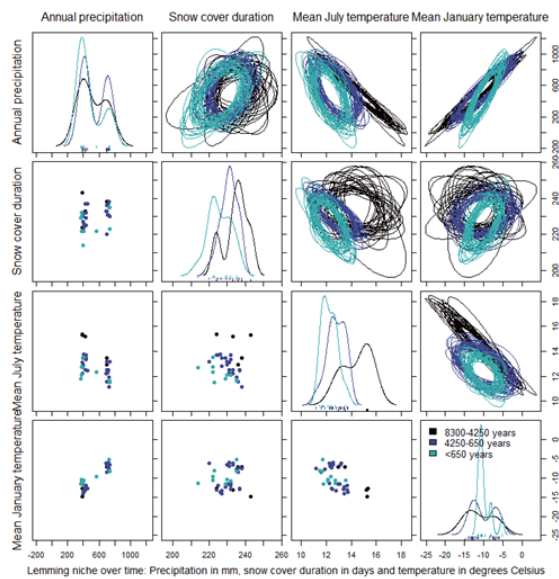
575 The pattern of change in climate niche dimension of Norwegian lemming is surprisingly
 576 similar to that of grey red-sided vole: the probability of finding the niche of Late Holocene or
 577 Historical period in the Middle Holocene was close to zero whereas there was less change in
 578 climate niche between Late Holocene and the Historical period (Supplementary Fig. 8.4a).
 579 The change in abiotic plant traits was similar to that of climate (Supplementary Fig. 8.4b),
 580 and also to that of grey red-sided vole (Supplementary Fig. 8.3b). Similarly, there was close
 581 to zero probability of finding the plant functional niche of the Middle Holocene in later
 582 periods, whereas the overlap was somewhat higher for plant abiotic traits and plant
 583 taxonomic diversity (Supplementary Fig. 8.4b-d). The probability of finding the niche from the
 584 Historical period in the Late Holocene was low for climate niches (~21%) but high for the

three others (60-80%) as well as the combination of all 15 niche dimensions (~81%,
Supplementary Fig. 8.4).

The abiotic and plant taxonomic diversity niche dimensions were largest in Late Holocene
and a similar trend was also seen for plant functional type niche dimensions (Supplementary
Fig. 8.1). This was also the overall pattern based on all 15 individual niche dimensions (Fig.
3c). Thus, similar to grey-sided vole, the niche decreased in the last 650 years whereas the
niche overlap with reindeer was high 88-94% (Supplementary Table S7).

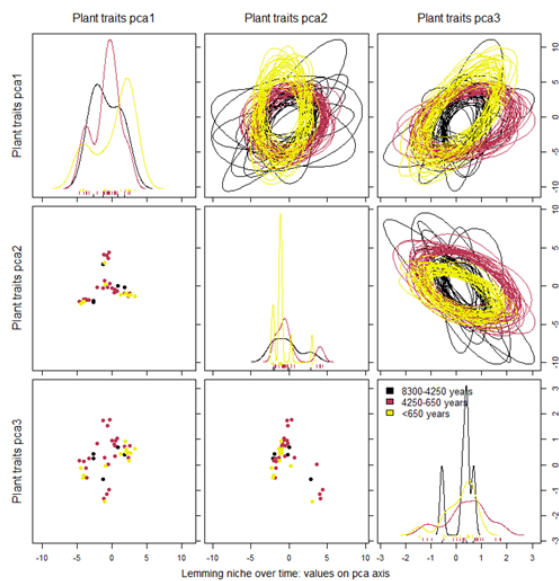
595

596a)



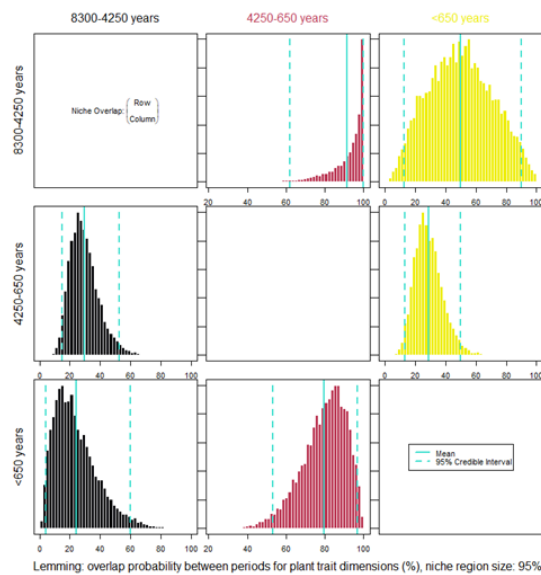
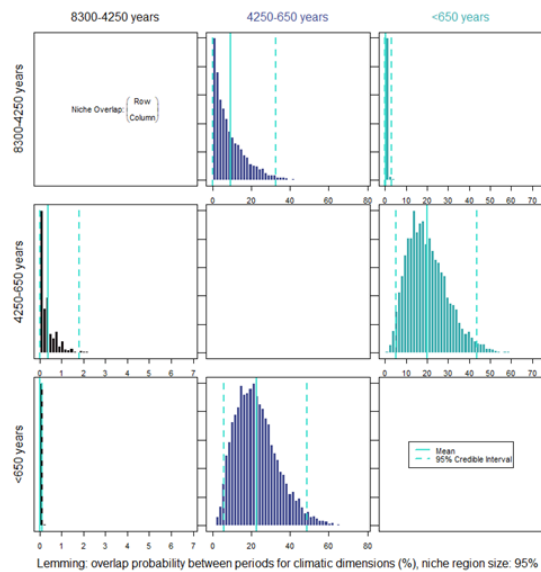
597

598b)

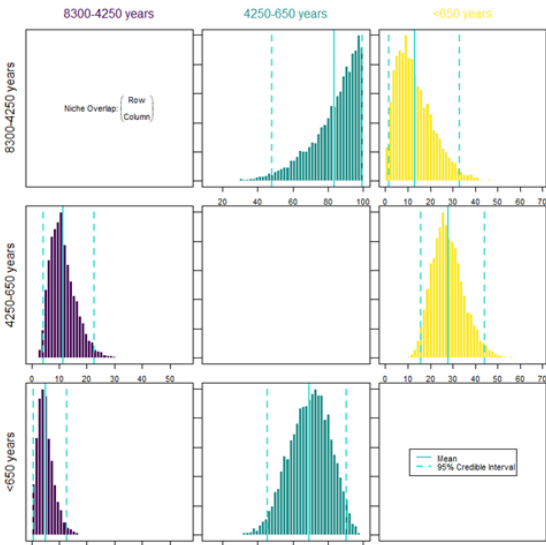
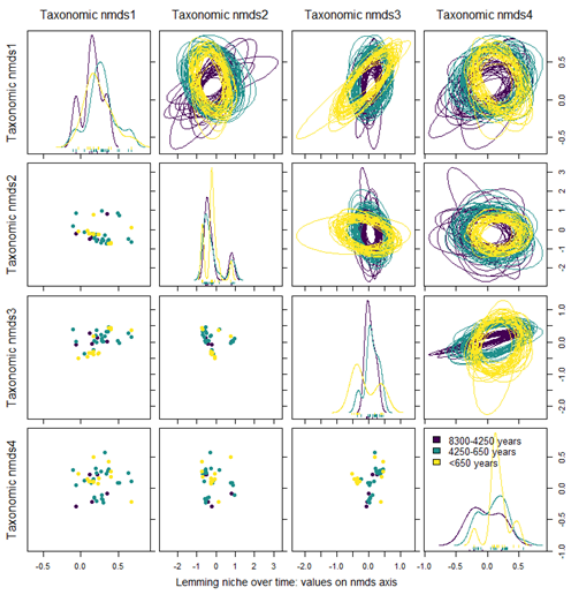


599

600

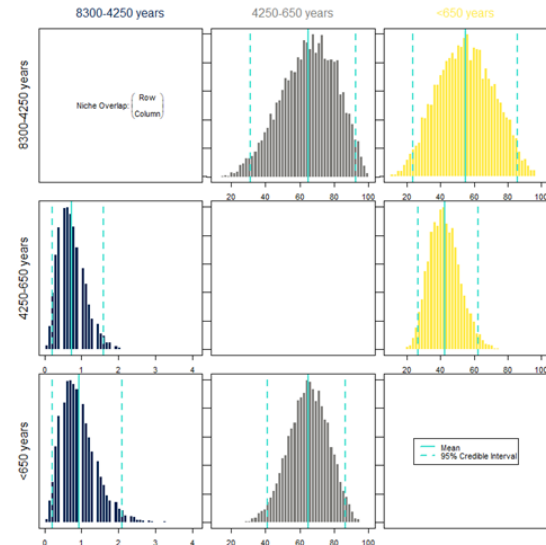
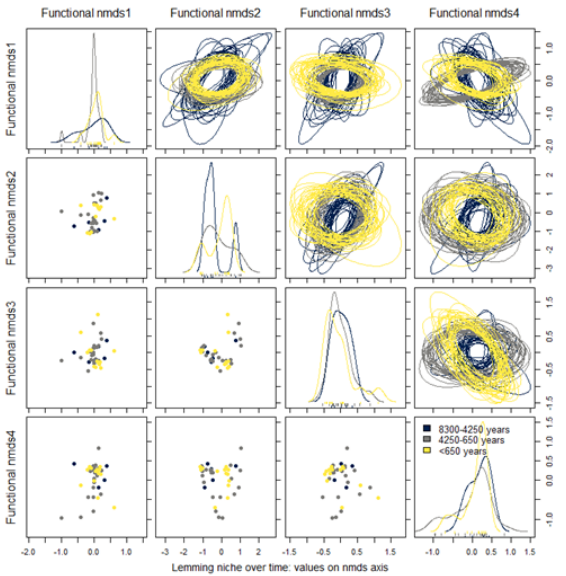


601c)



602

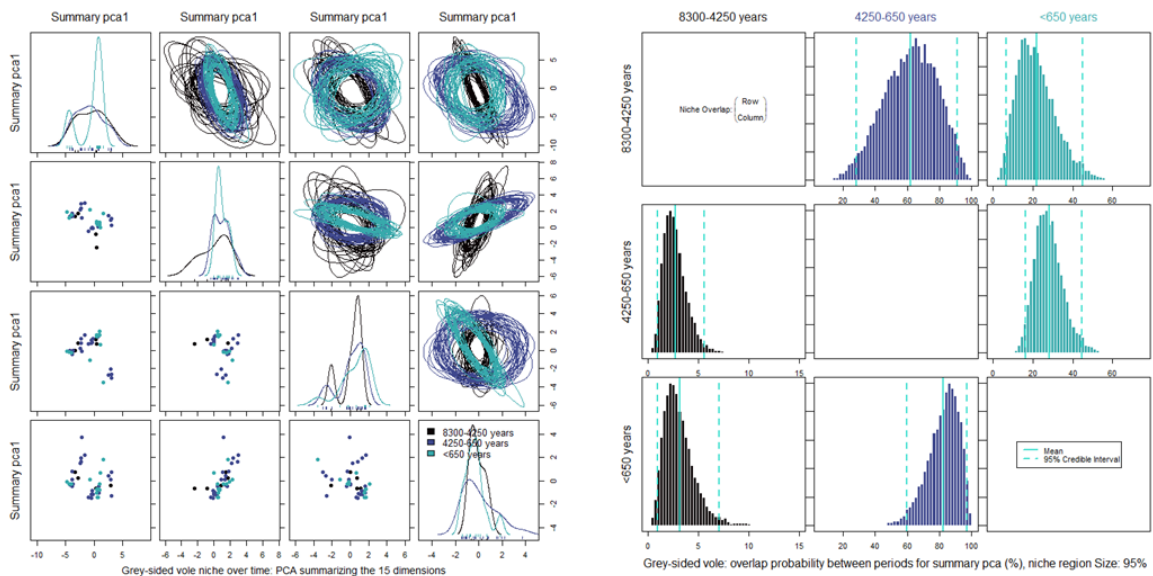
603d)



604

605

606 e)



607

608

609 **Supplementary Fig. 8.4. Niche stability over time for *Norwegian lemming* (*Lemmus lemmus*)**
 610 **based on a, climate niche dimension, b, climate and other abiotic conditions inferred from**
 611 **the traits of plants present, c, vegetation attributes with emphasis on taxonomic richness, d,**
 612 **vegetation attributes using with emphasis on abundance of plant functional composition, e,**
 613 **all 15 niche dimensions summarised using 3 PCA axis representing 70% of the variation.**
 614 **The bar plots represent the posterior probability that an individual from the time period**
 615 **indicated by the row will be found within the niche of the time period indicated by the column**
 616 **header.**

617

618 8.5 Willow ptarmigan niche

619 The single niche dimension that predicted the occurrence of willow ptarmigan best was plant
 620 functional group NMDS dimension 4 followed by plant taxonomic diversity NMDS dimension
 621 1 and abiotic traits PCA dimension 3. The set of niche dimensions that explained the
 622 occurrence of willow ptarmigan best was plant functional groups (40%) followed by plant
 623 taxonomic diversity (33%), whereas the climate (13%) and abiotic plant traits (14%) (Fig. 3).

624

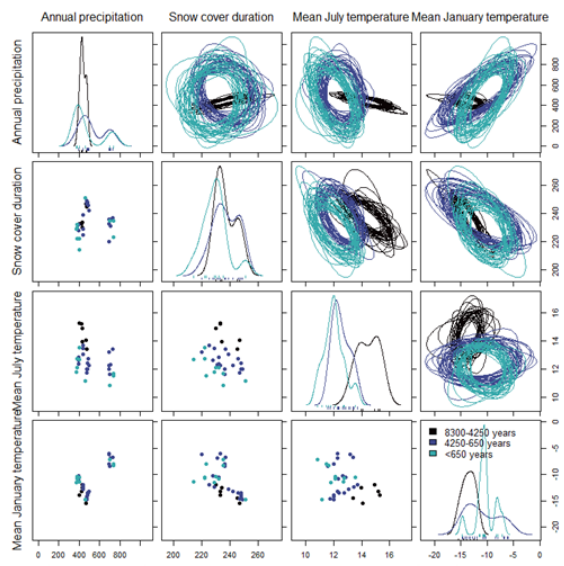
625 The probability of finding the Late Holocene and Historical niche dimensions in the Middle
 626 Holocene was <5% for all four sets of niche dimensions and also for the combination of the
 627 15 dimensions (Supplementary Fig. 8.5). For example, willow ptarmigan occurred at higher
 628 mean July temperature in the Early Holocene (12-16°C) than the Middle Holocene and
 629 Historical period (10-14°C). The probability of finding the niche dimensions of the last 650
 630 years in Late Holocene was somewhat higher for plant taxonomic diversity (~60%)
 631 compared to climate, abiotic traits, and plant functional groups (35-50%), and was low even
 632 for all 15 dimensions combined (mean ~36%). This is also reflected in the two-dimensional
 633 niche rover plots, which show low overlap among periods.

634

635

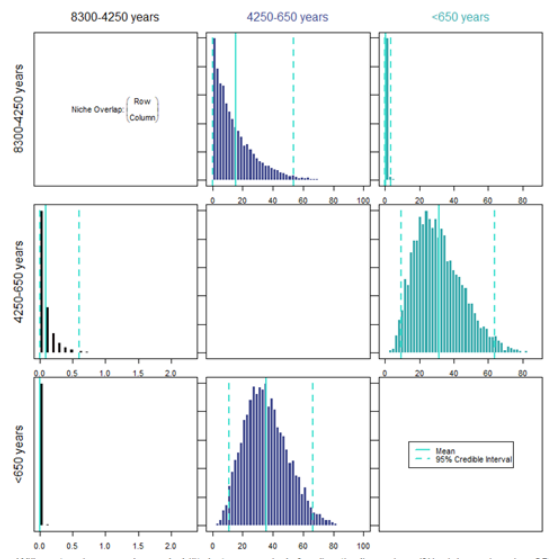
636

637a)



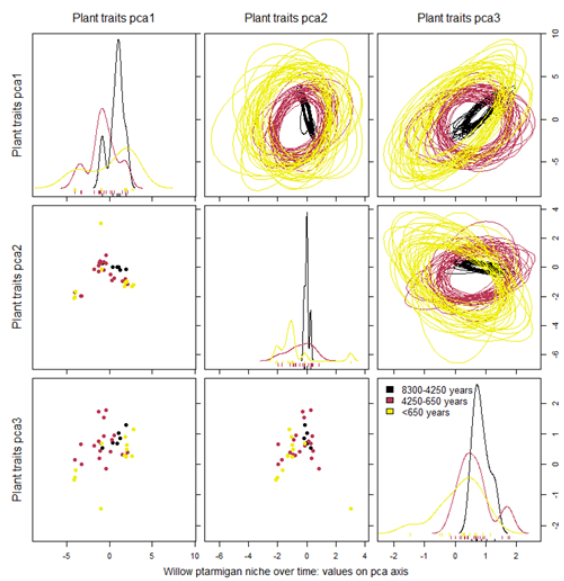
638

Willow ptarmigan niche over time: Precipitation in mm, snow cover duration in days and temperature in degrees Celsius



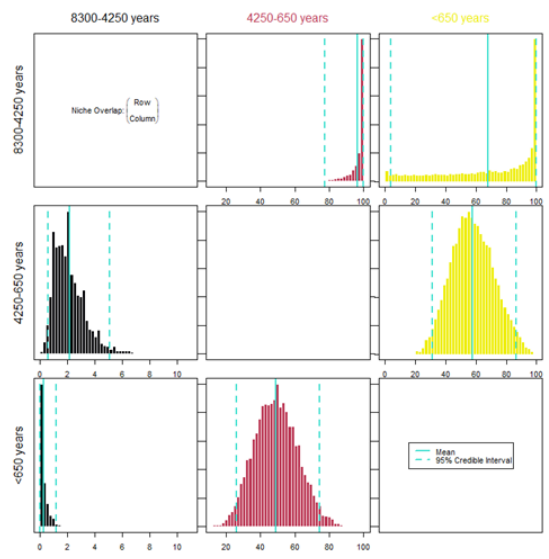
Willow ptarmigan: overlap probability between periods for climatic dimensions (%), niche region size: 95%

639b)



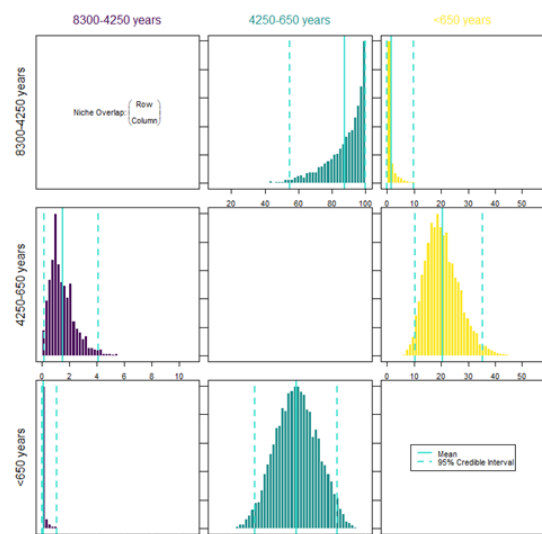
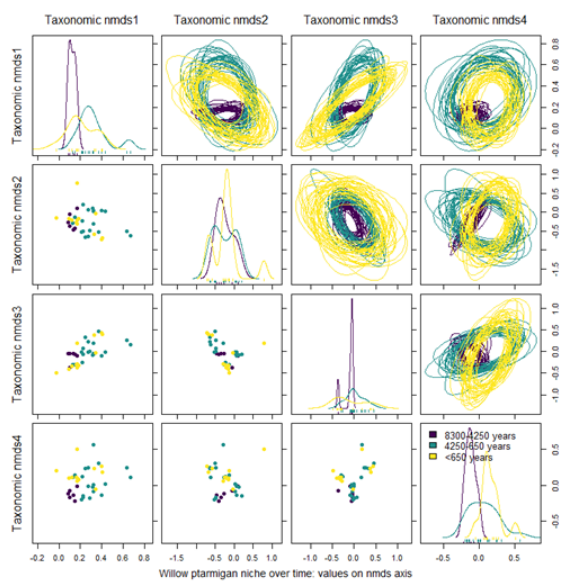
640

641



Willow ptarmigan: overlap probability between periods for plant trait dimensions (%), niche region size: 95%

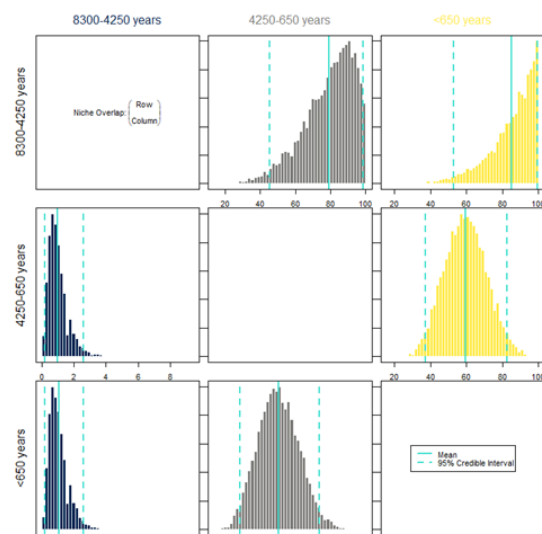
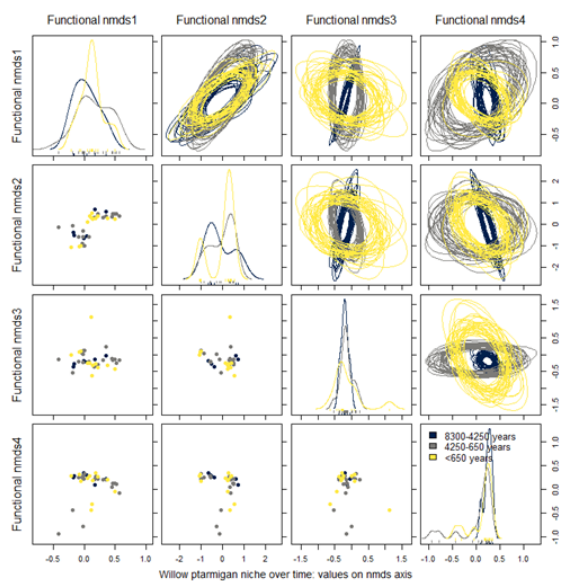
642c)



Willow ptarmigan: overlap probability between periods for plant taxonomy dimensions (%), niche region size: 95%

643

644d)

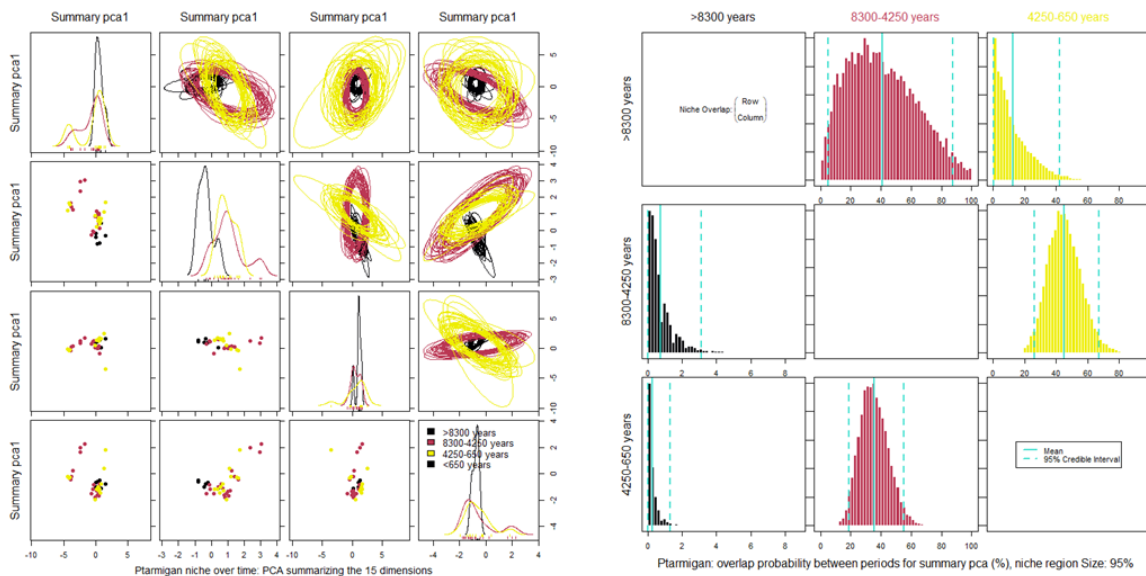


Willow ptarmigan: overlap probability between periods for plant functional group dimensions (%), niche region size: 95%

645

646

647e)



648

649

650 **Supplementary Fig. 8.5. Niche stability over time for *willow ptarmigan* (*Lagopus lagopus*)**
 651 **based on a, climate niche dimension, b, climate and other abiotic conditions inferred from**
 652 **the traits of plants present, c, vegetation attributes with emphasis on taxonomic richness, d,**
 653 **vegetation attributes using with emphasis on abundance of plant functional composition, e,**
 654 **all 15 niche dimensions summarised using 3 PCA axis representing 70% of the variation.**
 655 **The bar plots represent the posterior probability that an individual from the time period**
 656 **indicated by the row will be found within the niche of the time period indicated by the column**
 657 **header.**

658 8.6 Frog niche

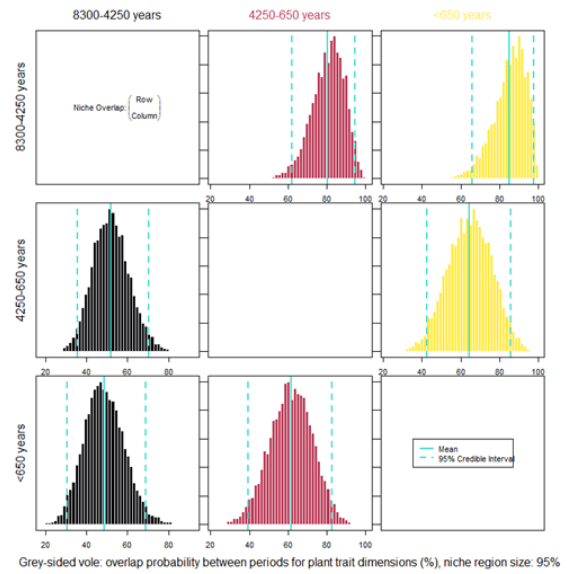
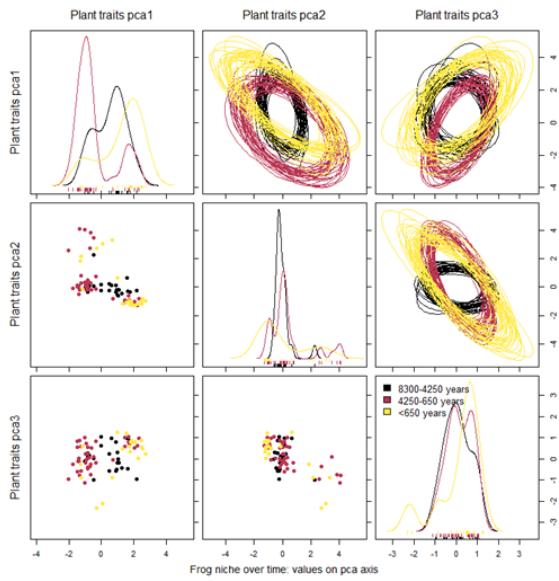
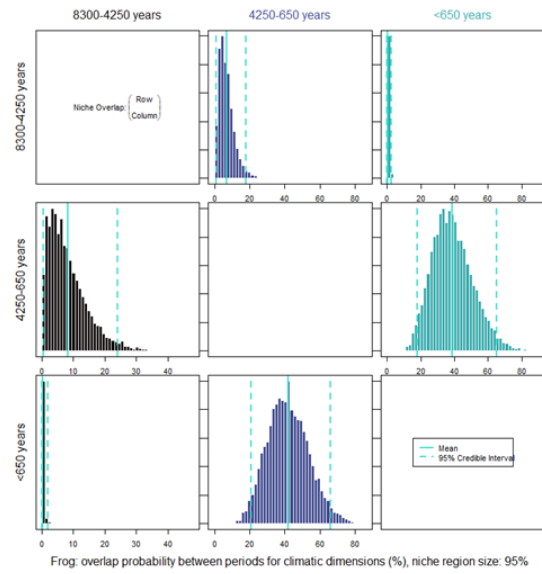
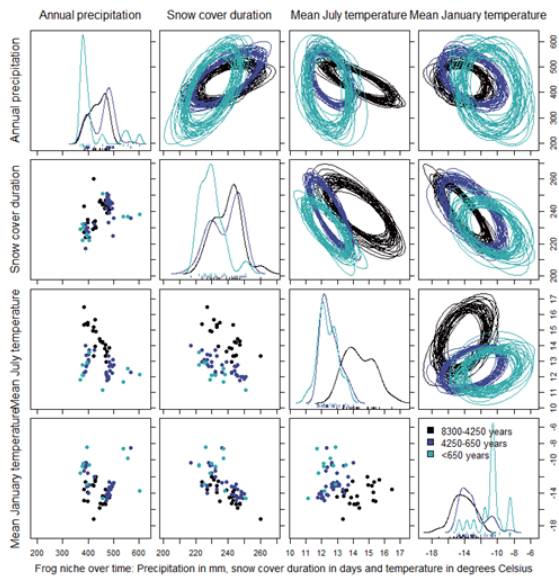
659 Like for reindeer, the single niche dimension that predicted frog occurrence best was annual
 660 precipitation, followed by taxonomic diversity NMDS dimension1. Other important predictors
 661 were taxonomic diversity NMDS dimension 3 and plant functional groups NMDS dimension 2
 662 (Fig. 3). Also, similar to reindeer, the set of niche dimensions that explain the occurrence of
 663 frog best was climate (39%) followed by plant taxonomic diversity (29%), whereas plant
 664 functional and plant abiotic traits explained 13% and 18%, respectively.

665

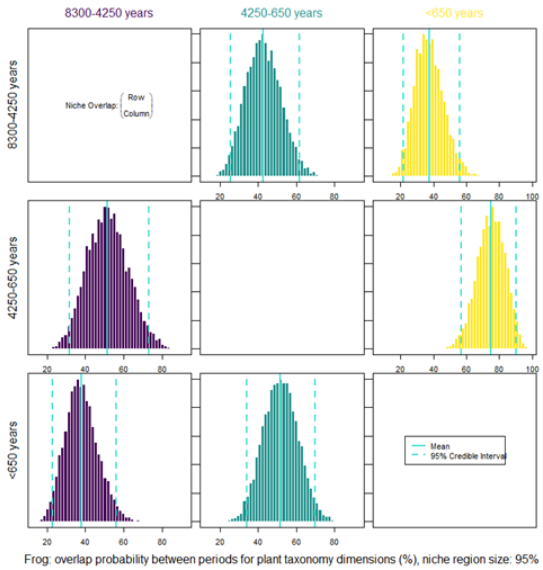
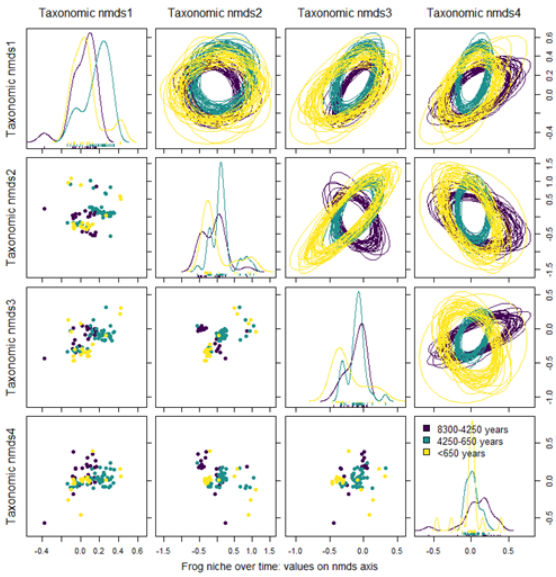
666 Except for the climatic niche (Supplementary Fig. 8.6a), there was more overlap between the
 667 Late Holocene and Middle Holocene niche for frog than for the terrestrial herbivores, with
 668 ~50% probability for abiotic plant traits, taxonomic plant diversity and plant functional groups
 669 and 40% for all 15 dimensions combined (Supplementary Fig. 8.6b-e).

670

671 As frog is both a carnivore and has exothermal regulation of body heat, it is not surprising
 672 that the climate niche dimensions predicted the occurrence best. Nevertheless, plant
 673 taxonomic diversity was the second most important niche dimension, probably reflecting
 674 characteristics of the habitat. For frog, we included the aquatic plants, and they may be good
 675 indicators of both biotic and abiotic conditions in the lake⁵¹. Thus, the importance of
 676 taxonomic diversity of plant may be due to aquatic plants reflecting habitat conditions
 677 adequate for the frog.



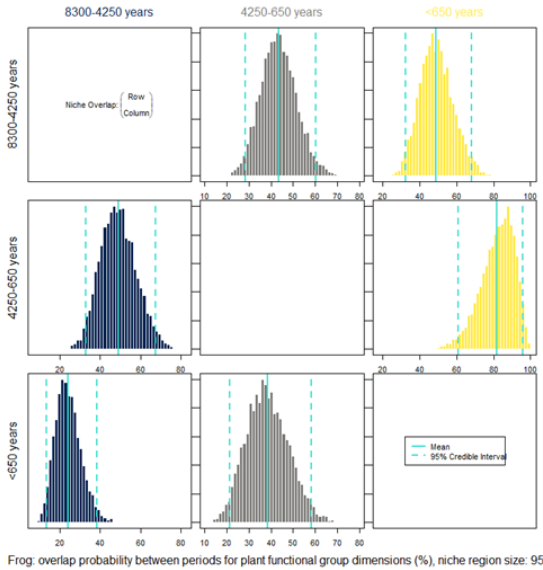
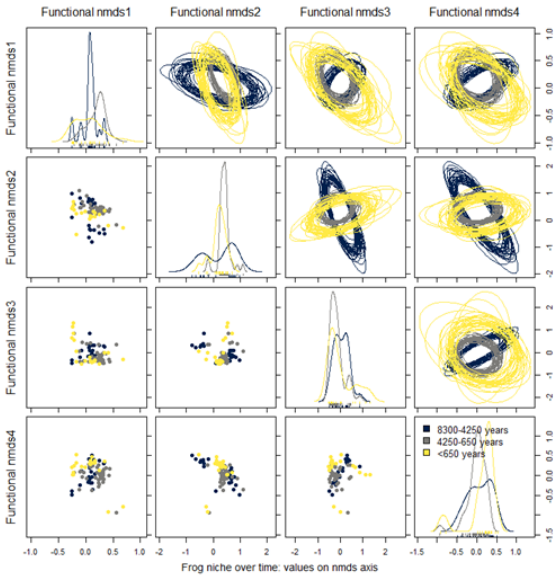
683c)



684

685

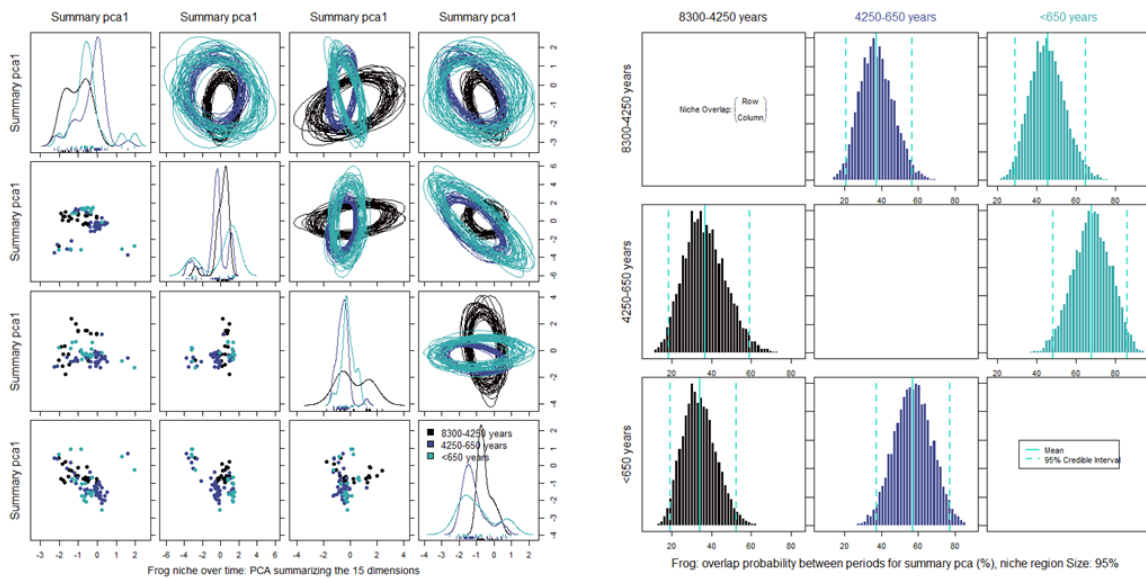
686d)



687

688

689e)



690

691 **Supplementary Fig. 8.6. Niche stability over time for *frog* (*Rana temporaria*) based on a,**
 692 **climate niche dimension, b, climate and other abiotic conditions inferred from the traits of**
 693 **plants present, c, vegetation attributes with emphasis on taxonomic richness, d, vegetation**
 694 **attributes using with emphasis on abundance of plant functional composition, e, all 15 niche**
 695 **dimensions summarised using 3 PCA axis representing 70% of the variation. The bar plots**
 696 **represent the posterior probability that an individual from the time period indicated by the row**
 697 **will be found within the niche of the time period indicated by the column header.**

698 9 Change in environmental space

699 The environmental space changed over time for all four sets of niche dimensions
700 (Supplementary Fig. 9.1). The climate (mean annual temperature, temperature variability,
701 annual precipitation and mean July temperature) was unstable in the Early and Middle
702 Holocene, and there was also most uncertainty in estimating the climate environmental
703 space in these periods. Climate variability, especially annual precipitation and mean July
704 temperature, narrowed from the Middle Holocene until present. There was less than 50%
705 overlap between the last 650 years and the periods prior to 4.25 ka. The environmental
706 space estimated from abiotic plant traits also showed the largest change between Early and
707 Middle Holocene compared to Late and Historical period, but the overlap in environmental
708 space was higher (70-90% for most comparisons). Thus, in terms of abiotic factors, there
709 was a relatively larger change in climate than in other abiotic factors such as nitrogen,
710 phosphate, pH and light.

711

712 The taxonomic richness showed the largest change from a wider taxonomic niche in the
713 Early Holocene that only showed a mean overlap of <30% with later periods. As the peak
714 arrival of plants was around 12 ka²⁸, the change is likely not due to limited data in the early
715 period but rather represents a shift from a mixture of tundra and boreal habitats to boreal
716 vegetation dominating most sites. In contrast, the plant functional dimensions, which
717 emphasise the abundance of different plant groups, showed higher overlap in niche
718 dimensions and mean probability of overlap of 50% or higher in all cases, probably reflecting
719 that different plant species may fill the same function over time.

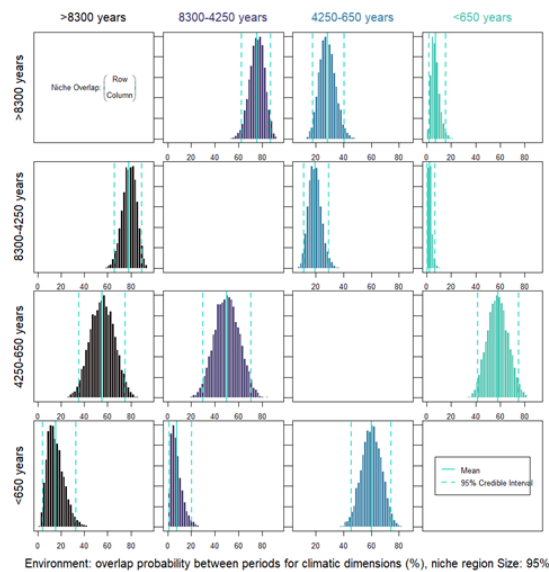
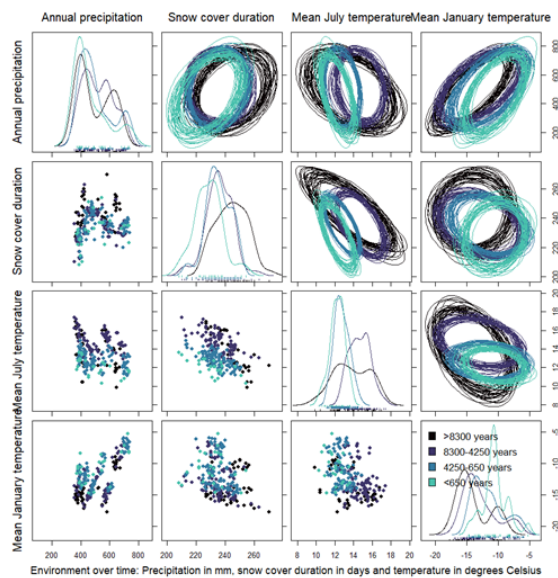
720

721 When evaluating all 15 niche dimensions, the largest change was between the two earliest
722 time periods compared to the two recent periods (Supplementary Fig. 9.1e). Especially, the
723 probability of finding the environmental space of the Early Holocene was low for the Late
724 Holocene (~38%) and Historical period (~20%), whereas the probability of finding the
725 environmental space of the Middle Holocene in Late Holocene and the Historical period was
726 somewhat higher (40-60%).

727

728

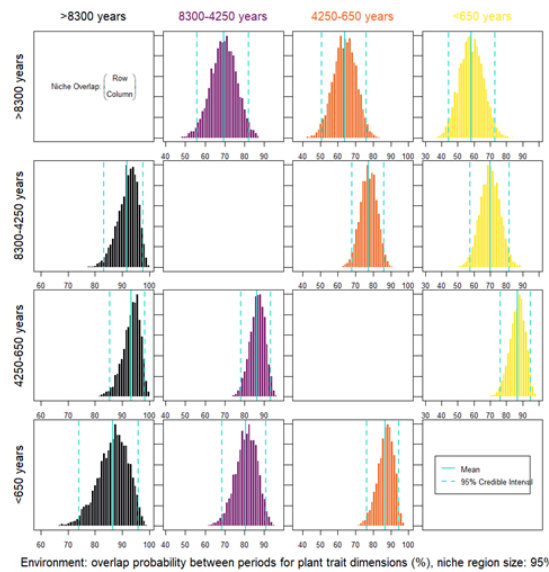
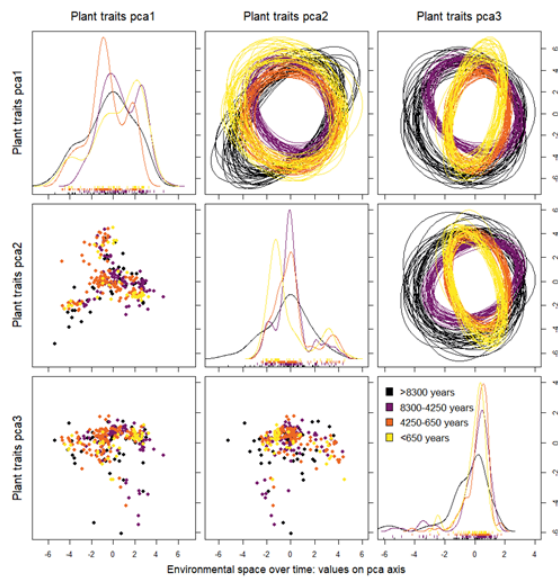
729a)



730

731

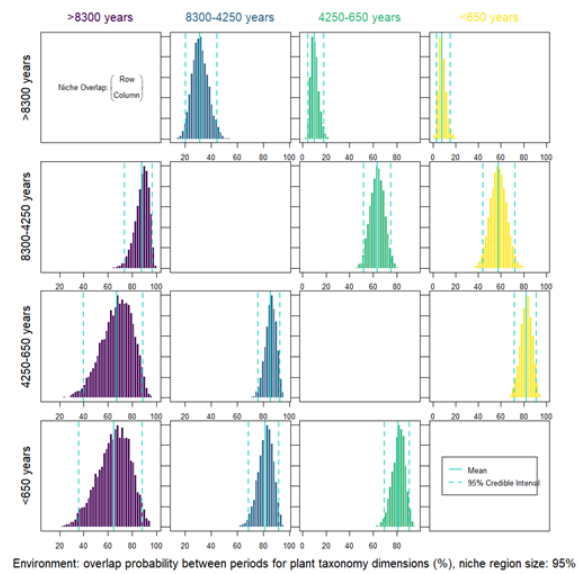
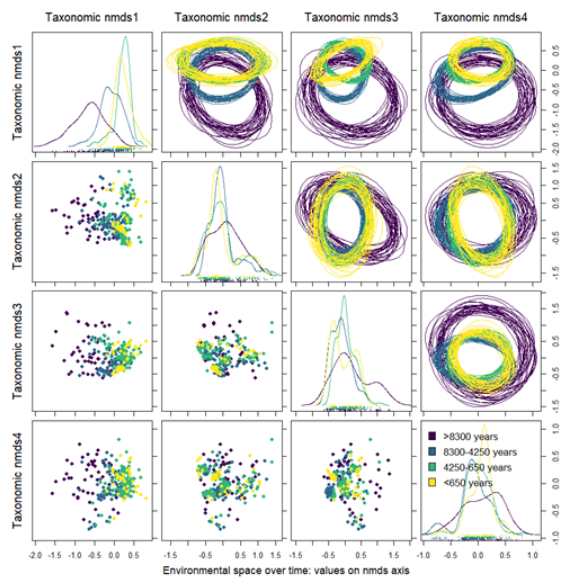
732b)



733

734

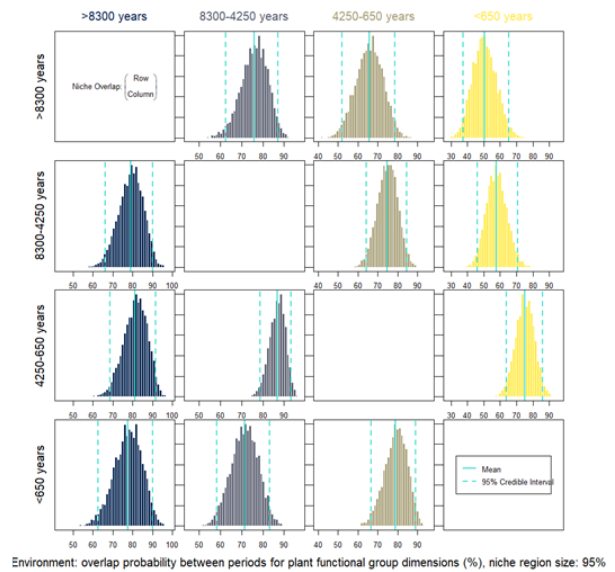
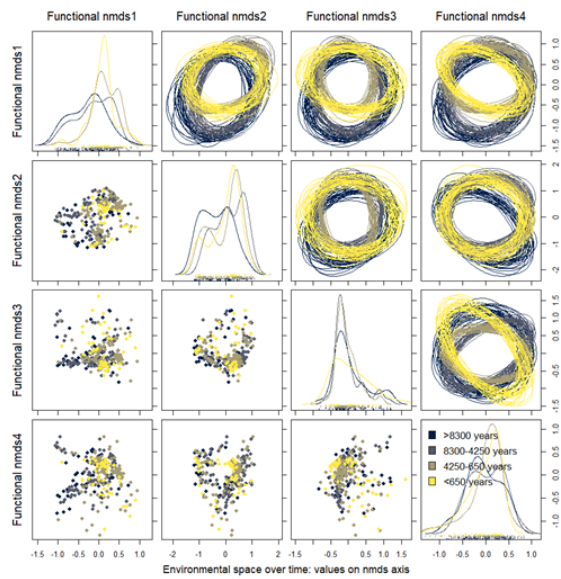
735c)



736

737

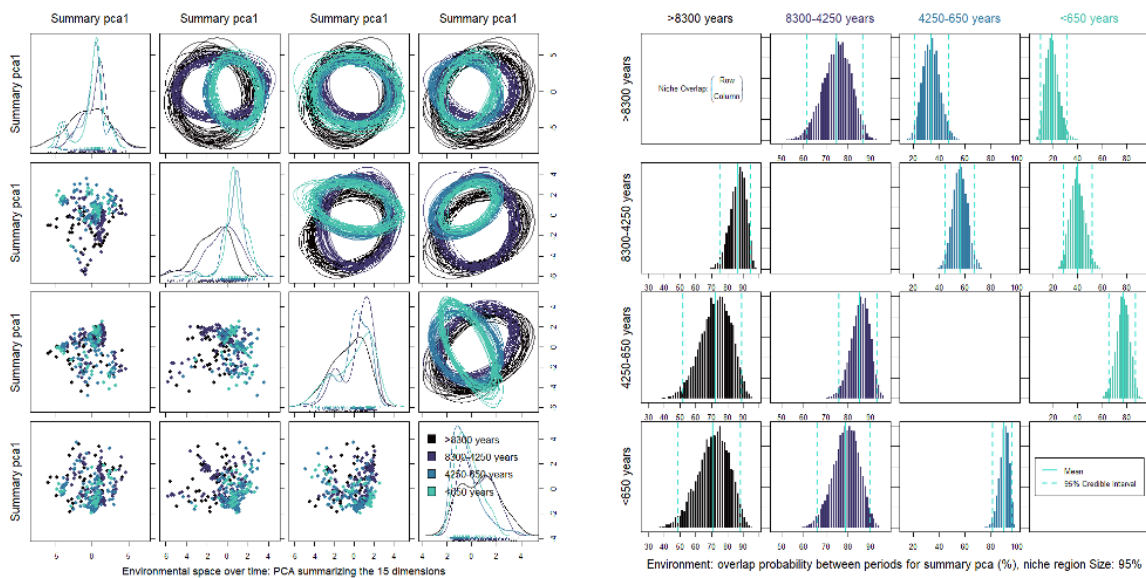
738d)



739

740

741 e)



742

743

744

745 *Supplementary Fig. 9.1. Changes in **environmental space** based on **a**, climate niche*
 746 *dimension, **b**, climate and other abiotic conditions inferred from the traits of plants present, **c**,*
 747 *vegetation attributes with emphasis on taxonomic richness, **d**, vegetation attributes using*
 748 *with emphasis on abundance of plant functional composition, **e**, all 15 niche dimensions*
 749 *summarised using 3 PCA axis representing 70% of the variation. The bar plots represent the*
 750 *posterior probability that ecological conditions from the time period indicated by the row will*
 751 *be found within the time period indicated by the column header.*

752

753 10 Niche overlap among key herbivores

754 For the Middle Holocene, there was a high probability to find willow ptarmigan within the
 755 niche of reindeer for the four categories of niche dimensions as well as the combination of all
 756 (68-90%, Supplementary Fig. 10.1). There was also a medium high probability to find
 757 grey-sided vole in the abiotic niche and functional diversity niche of lemming and reindeer
 758 (50-60%), whereas the overlap was otherwise low.

759 Increasing niche overlap was seen in the Late Holocene where the probability of finding any
 760 of the four herbivores within the niche of each of the others were 60-80% for most niche
 761 dimensions, and ~90% for finding any of the three others within the abiotic traits and plant
 762 taxonomic dimension niche of Norwegian lemming (Supplementary Fig. 10.2). When
 763 combining all 15 niche dimensions, there was 70-83% chance to find any of the three other
 764 herbivores within the niche of willow ptarmigan, 50-60% to find them within the niche of
 765 Norwegian lemming, 58-68% to find them within the niche of reindeer and 80-85% mean
 766 probability to find them within the niche of grey-sided vole.

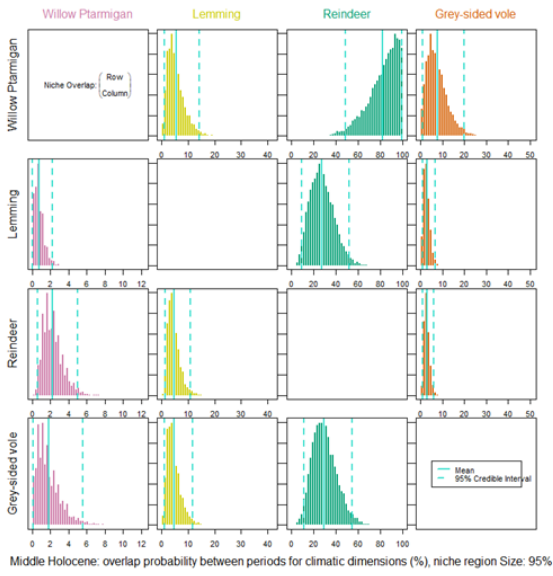
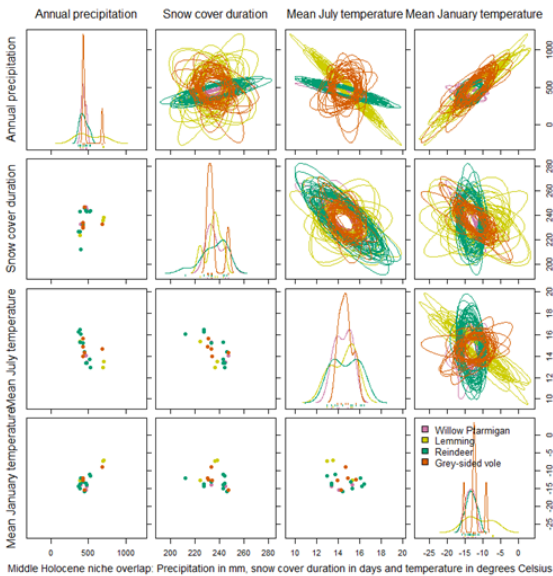
767 For the Historical period (Supplementary Fig. 10.3), the niche overlap was more complex
 768 than in the Late Holocene. The probability to find any of the herbivores within the climatic
 769 niche of others was still 60-80% except for a lower probability to find willow ptarmigan and

770 reindeer within the climatic niche of Norwegian lemming, and a ~90% probability to find any
771 of the smaller herbivores within the climatic niche of reindeer. For abiotic traits, there was a
772 high probability (91-95%) of finding the smaller herbivores within the niche of reindeer,
773 whereas the probability to find reindeer within the smaller niches of Norwegian lemming and
774 grey-sided voles was only 45% and 20%, respectively. For plant taxonomic diversity, the
775 probability of finding reindeer within the niche of any of the smaller herbivores was low
776 (10-25%), in contrast to 80-97% probability of finding the smaller herbivores within the
777 taxonomic niche dimension of reindeer. There was also a low probability of finding willow
778 ptarmigan and Norwegian lemming within the plant taxonomic niche of grey-sided vole,
779 reflecting the overall narrower niche width of grey-sided vole. In contrast, there was a high
780 probability of finding willow ptarmigan and grey-sided vole within the taxonomic niche of
781 lemming. For plant functional groups, there was medium to high overlap among all
782 herbivores (50-92%). When combining all 15 niche dimensions, the probability of finding any
783 of the herbivores within the niche of the smaller herbivores was slightly lower than in the
784 Late Holocene, whereas the probability of finding the smaller herbivores within the niche of
785 reindeer increased from 58-68% in the Late Holocene to 87-92% in the Historical period.

786

787

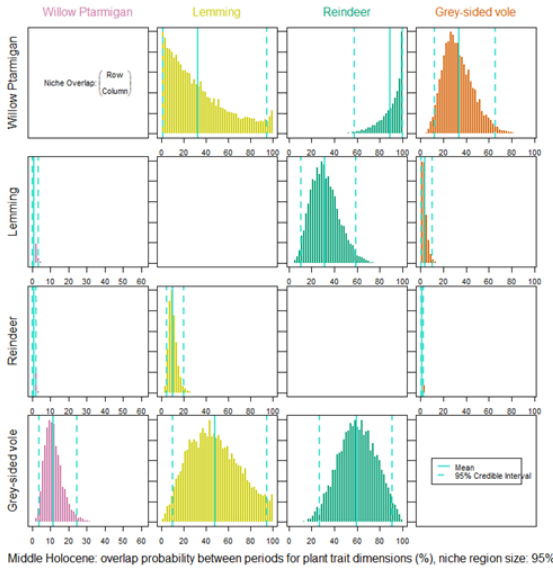
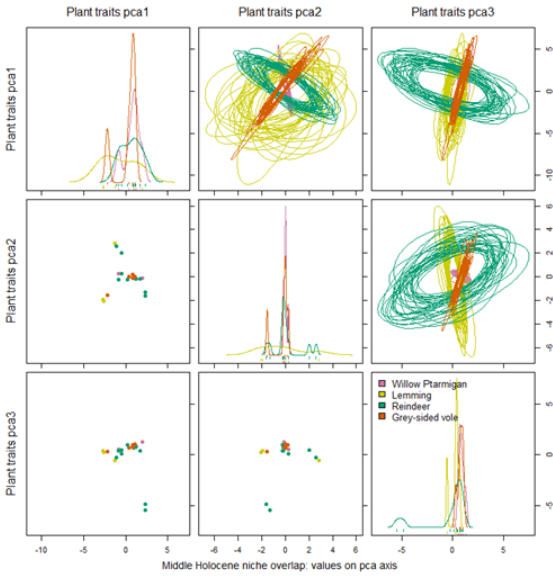
788a)



789

790

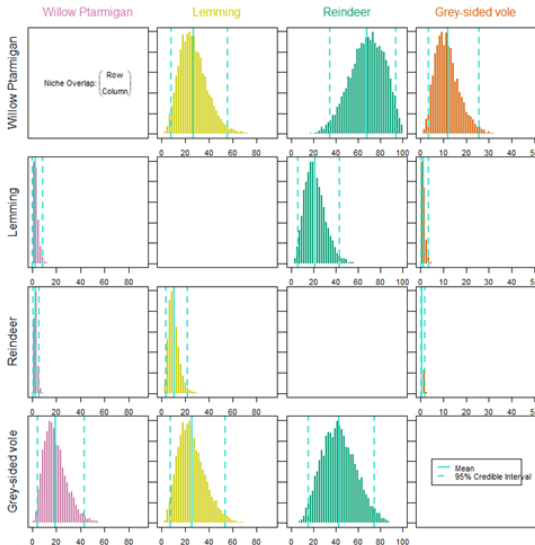
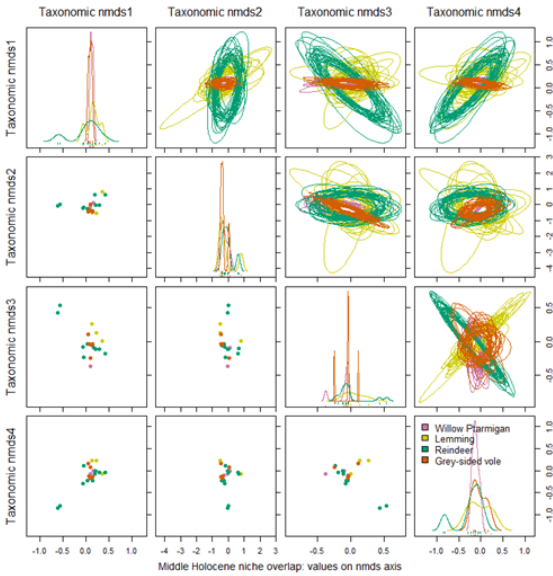
791b)



792

793

794c)

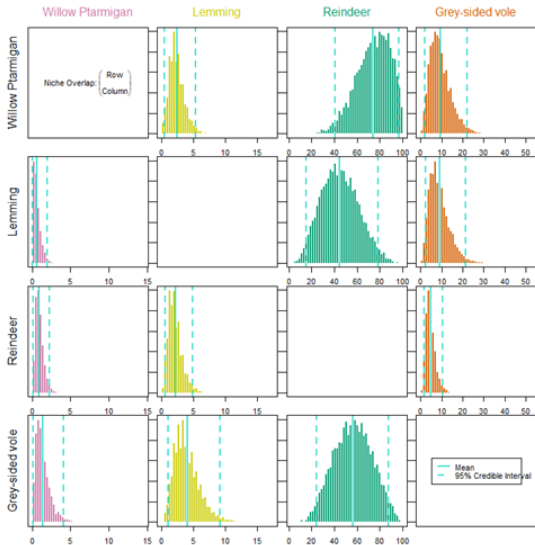
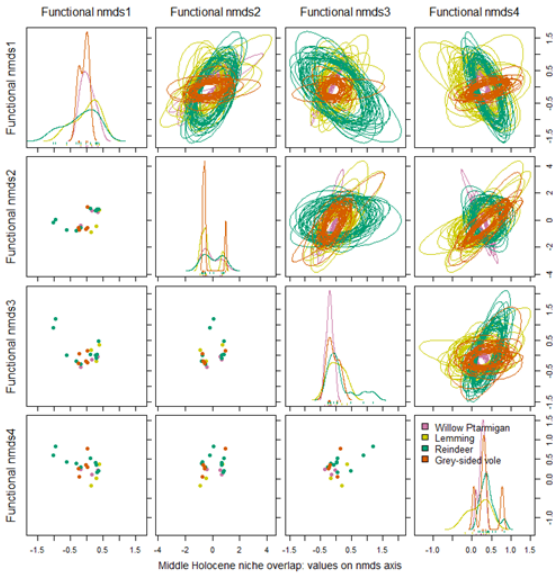


Middle Holocene: overlap probability between periods for plant taxonomy dimensions (%), niche region size: 95%

795

796

797d)

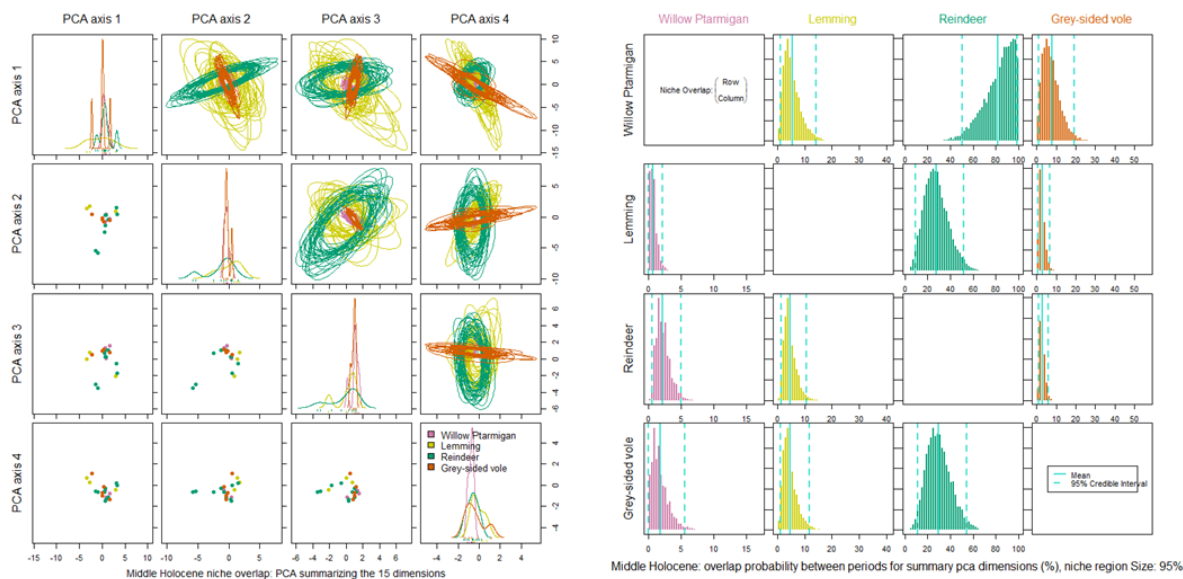


iddle Holocene: overlap probability between periods for plant functional group dimensions (%), niche region size: 95%

798

799

800e)



801

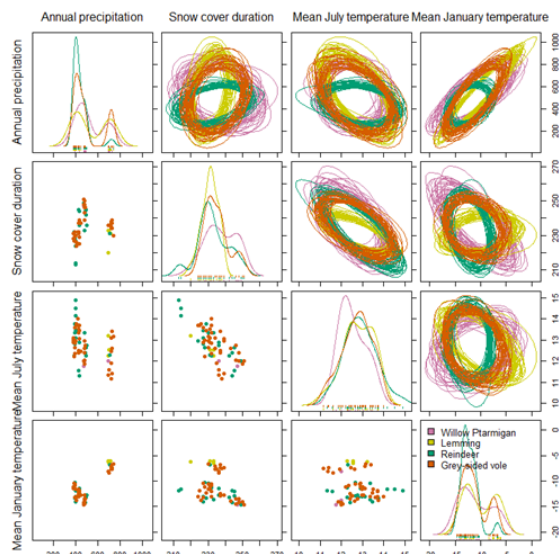
802

803 **Supplementary Fig. 10.1. Niche overlap among key herbivores during Middle Holocene**
804 **based on a, climate niche dimension, b, climate and other abiotic conditions inferred from**
805 **the traits of plants present, c, vegetation attributes with emphasis on taxonomic richness, d,**
806 **vegetation attributes using with emphasis on abundance of plant functional composition, e,**
807 **all 15 niche dimensions summarised using 3 PCA axis representing 70% of the variation.**
808 **The bar plots represent the posterior probability that a species indicated by the row will be**
809 **found within the niche of the species indicated by the column header.**

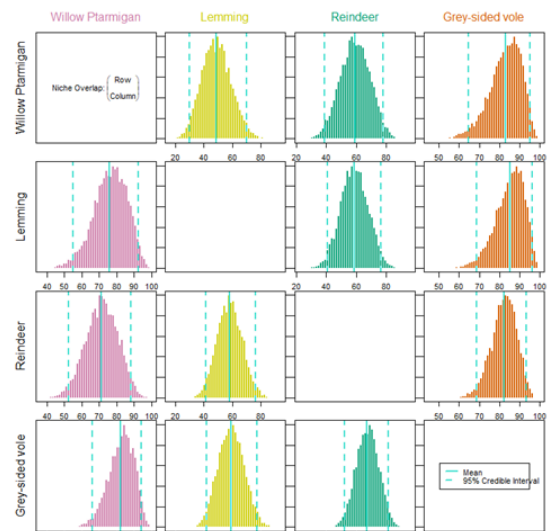
810

811

812a)



Late Holocene niche overlap: Precipitation in mm, snow cover duration in days and temperature in degrees Celsius

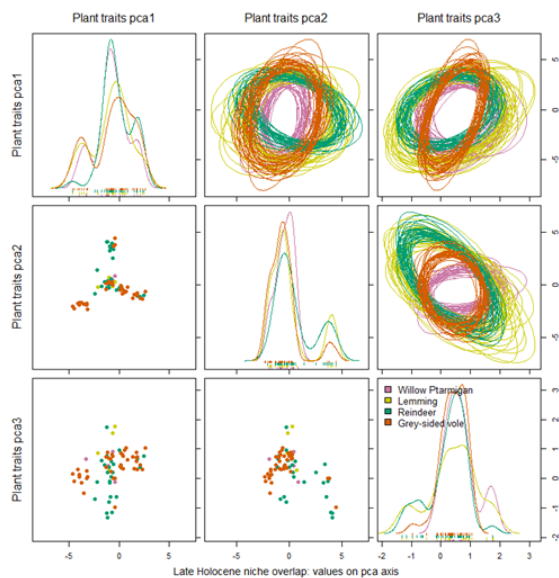


Late Holocene: overlap probability between periods for climatic dimensions (%), niche region size: 95%

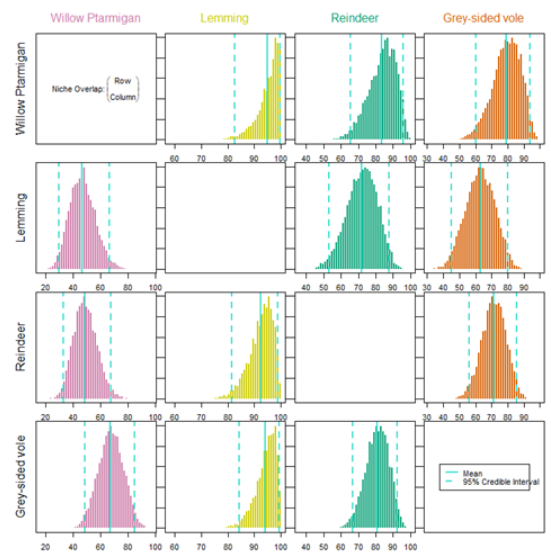
813

814

815b)



Late Holocene niche overlap: values on pca axis

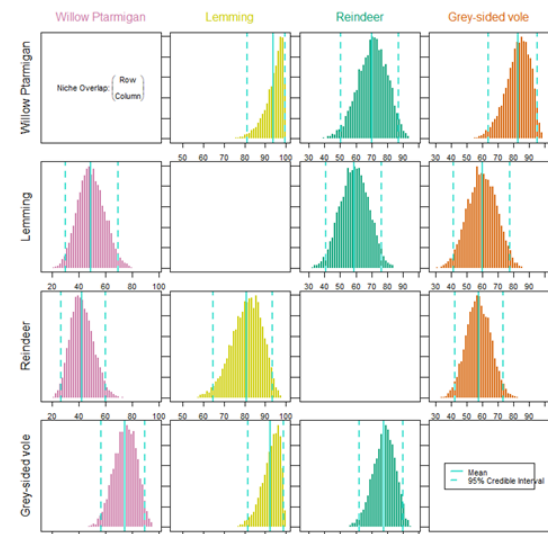
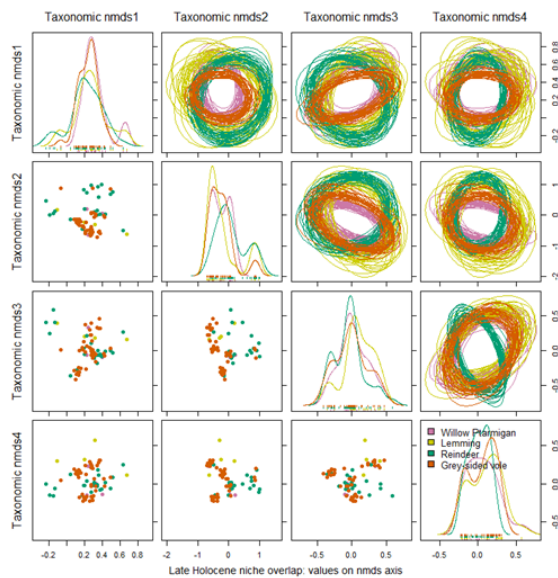


Late Holocene: overlap probability between periods for plant trait dimensions (%), niche region size: 95%

816

817

818c)

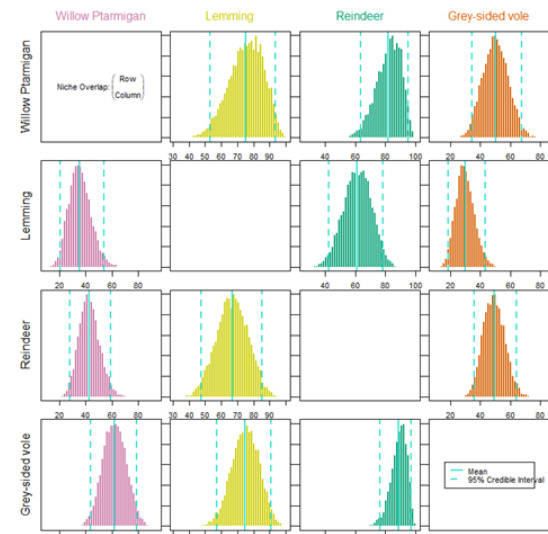
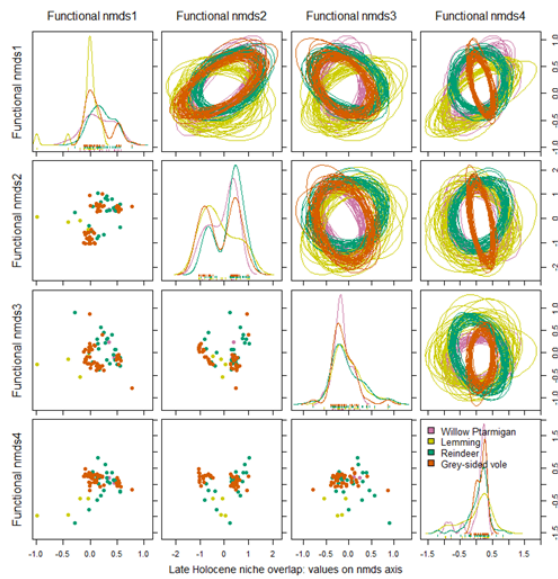


Late Holocene: overlap probability between periods for plant taxonomy dimensions (%), niche region size: 95%

819

820

821d)

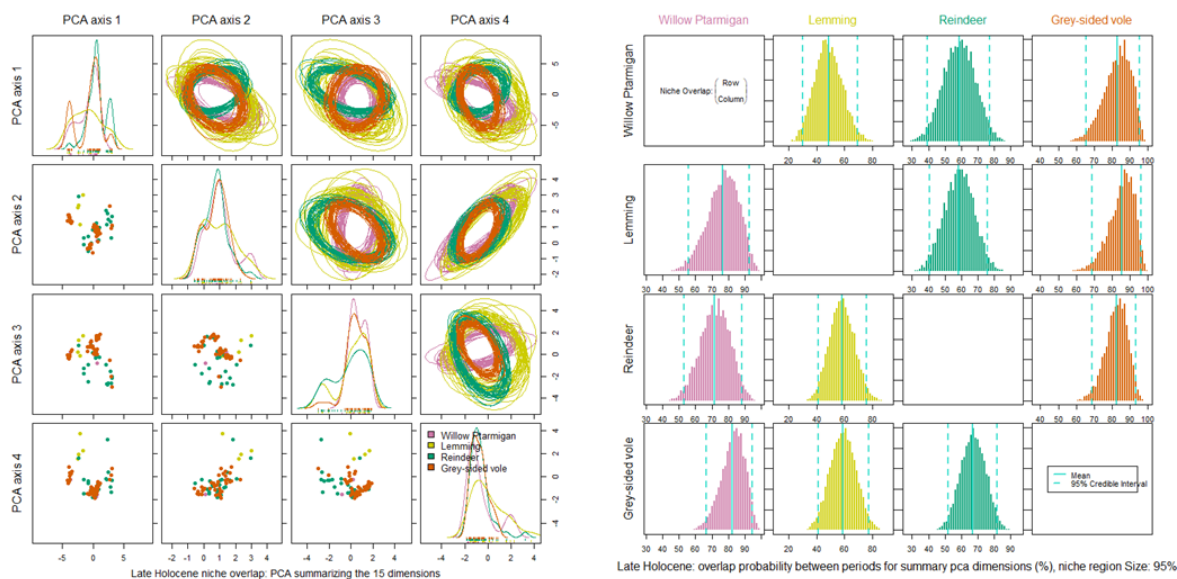


Late Holocene: overlap probability between periods for plant functional group dimensions (%), niche region size: 95%

822

823

824e)



825

826

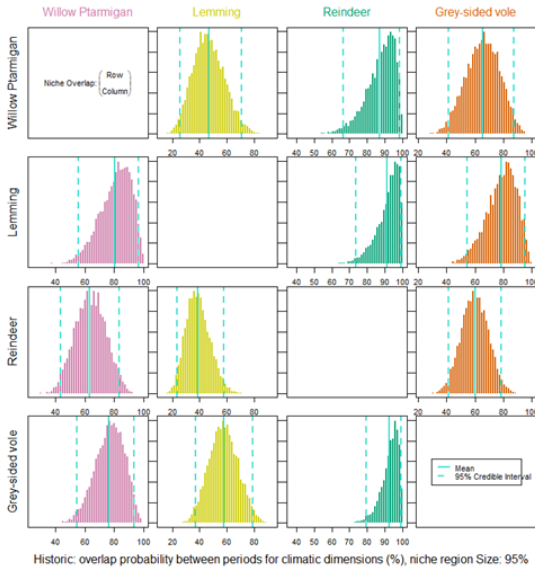
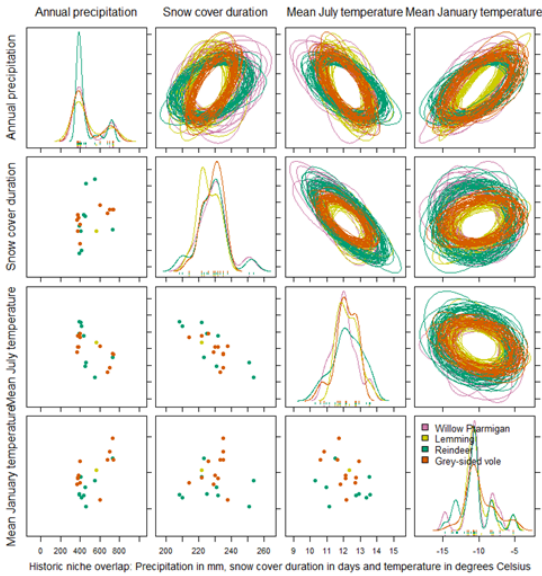
827 **Supplementary Fig. 10.2. Niche overlap among key herbivores during Late Holocene** based
828 on **a**, climate niche dimension, **b**, climate and other abiotic conditions inferred from the traits
829 of plants present, **c**, vegetation attributes with emphasis on taxonomic richness, **d**,
830 vegetation attributes using with emphasis on abundance of plant functional composition, **e**,
831 all 15 niche dimensions summarised using 3 PCA axis representing 70% of the variation.
832 The bar plots represent the posterior probability that a species indicated by the row will be
833 found within the niche of the species indicated by the column header.

834

835

836

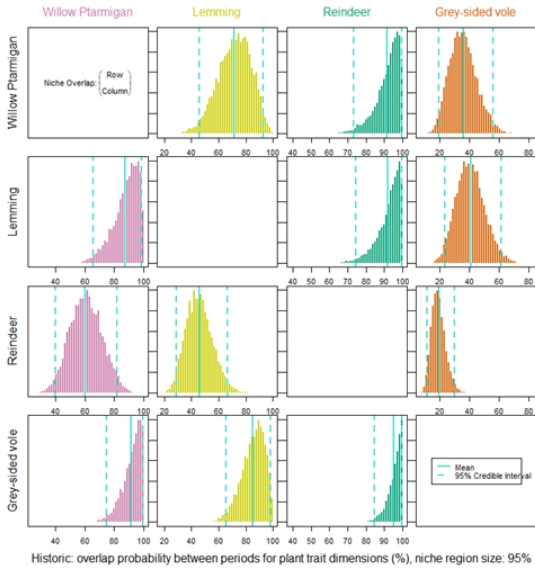
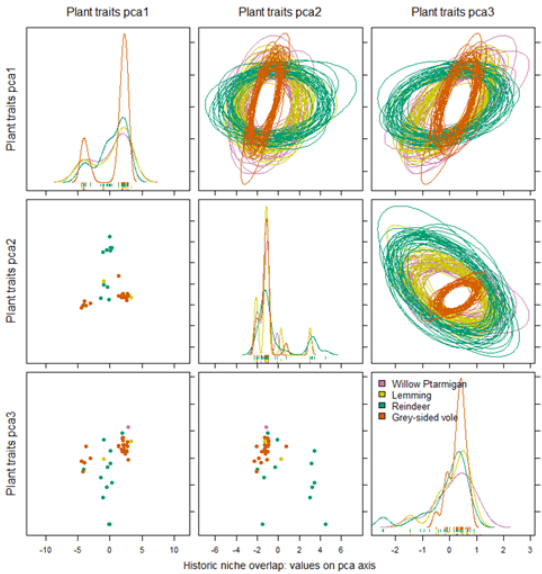
837a)



838

839

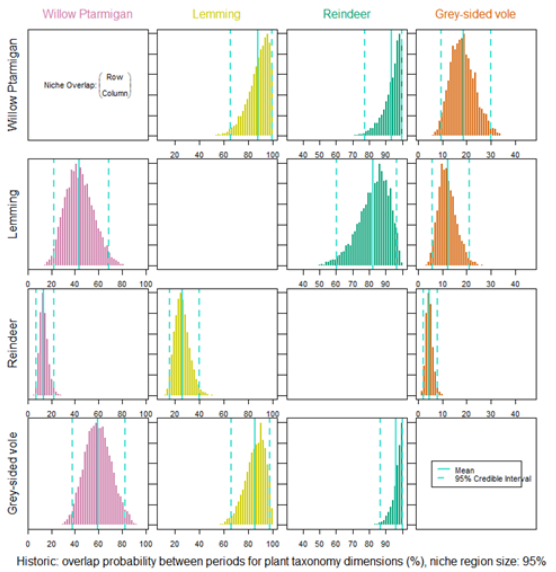
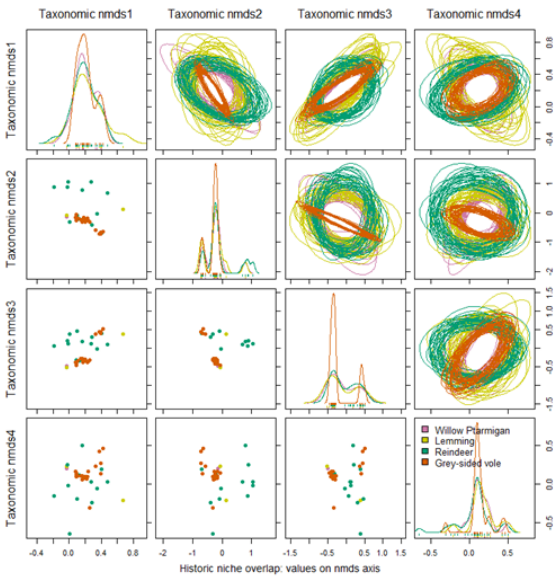
840b)



841

842

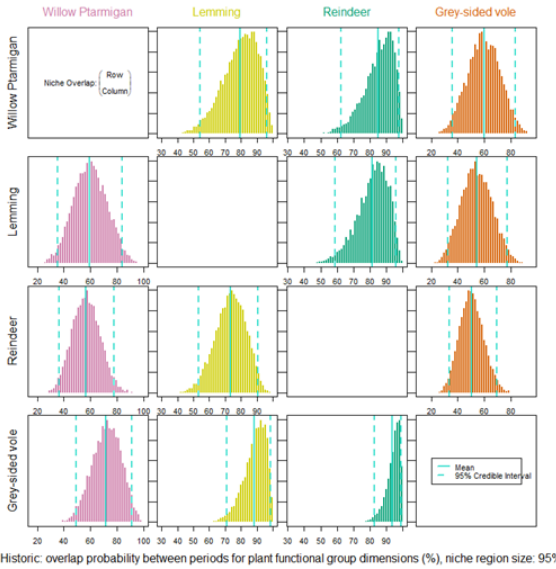
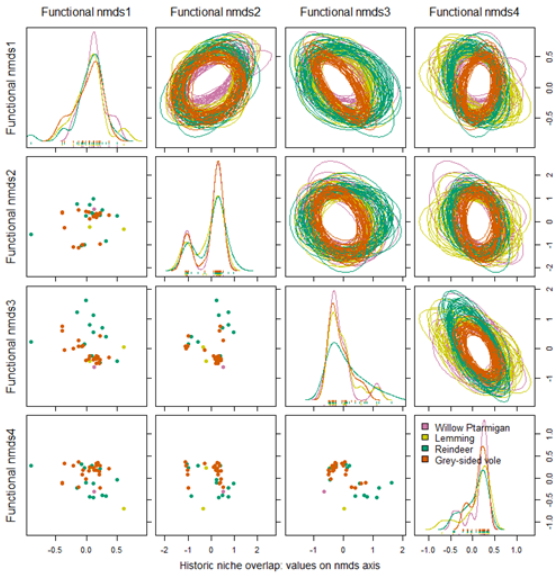
843 c)



844

845

846 d)

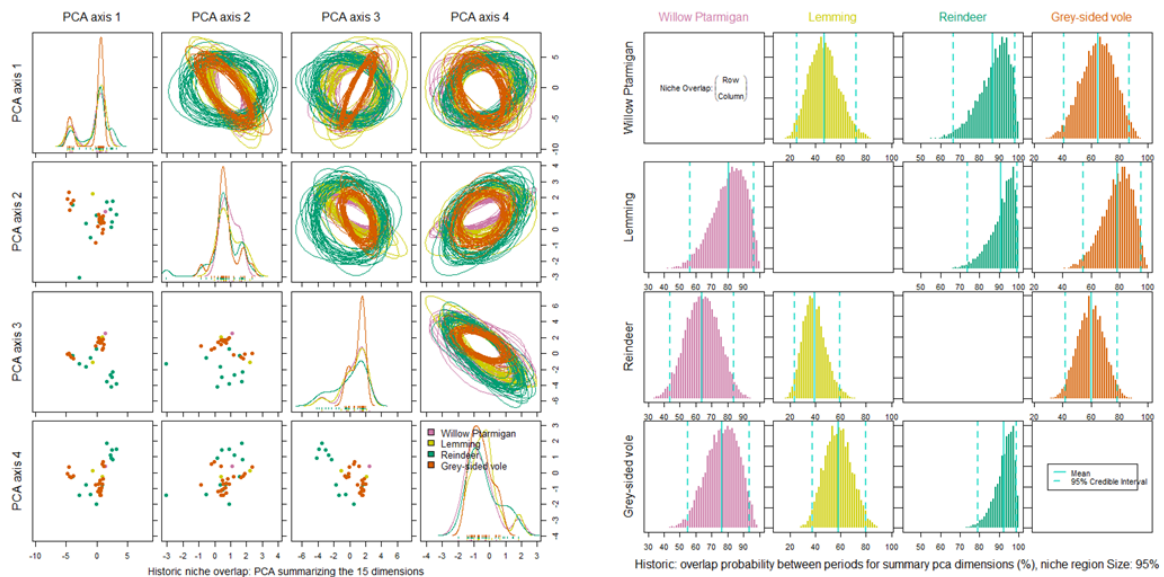


847

848

849

850e)



851

852

853 **Supplementary Fig. 10.3. Niche overlap among key herbivores during Historical time (last**
 854 **650 years) based on a, climate niche dimension, b, climate and other abiotic conditions**
 855 **inferred from the traits of plants present, c, vegetation attributes with emphasis on taxonomic**
 856 **richness, d, vegetation attributes using with emphasis on abundance of plant functional**
 857 **composition, e, all 15 niche dimensions summarised using 3 PCA axis representing 70% of**
 858 **the variation. The bar plots represent the posterior probability that a species indicated by the**
 859 **row will be found within the niche of the species indicated by the column header.**

860

861

862 11 Lemming and tundra vole 16S reference sequences

863 We supplemented our EMBL-derived animal 16S reference library with three sequences
 864 from Norwegian lemming (*Lemmus lemmus*), Siberian lemming (*Lemmus sibiricus*), and
 865 tundra vole (*Microtus oeconomus*). The Norwegian lemming sequence was previously
 866 presented in ⁵², whereas the Siberian lemming sequence was provided by David Díez del
 867 Molino and Love Dalén (Swedish Museum of Natural History/Stockholm University,
 868 Sweden).

869

870 Although a genome assembly is available for the tundra vole (NCBI RefSeq
 871 GCA_007455595.1), the mitochondrial genome is not included. We therefore searched the
 872 underlying raw sequence data used to generate this genome (NCBI BioProject
 873 PRJNA551187) to extract a tundra vole 16S sequence for use in the reference library. To do
 874 this, we grep-searched the raw sequence data for both the forward and reverse 16S P007
 875 primer sequences ⁵³ in both the forward and reverse complement orientations (for four
 876 searches in total). We then compiled the recovered reads and converted them to the same
 877 orientation. The reads were converted to FASTA format, aligned using the clustal algorithm ⁵⁴
 878 in SeaView v4 ⁵⁵, and a 90% majority-rule consensus sequence was generated in SeaView.
 879 Due to the short length of the barcode, this strategy allowed for the complete reconstruction

880 of the targeted 16S region. The consensus sequence was then trimmed to the barcode
881 coordinates and the reconstructed barcode underwent an NCBI BLAST search against
882 PRJNA551187, which had multiple 100% matches. Since this reference sequence was
883 generated, a further three tundra vole 16S sequences are now available on NCBI Genbank
884 (ON165388, ON165345, ON165278), which altogether have seven variable positions, of
885 which only one is unique to the sequence presented here. Altogether, these checks
886 demonstrate the validity of the reconstructed tundra vole 16S sequence used in our
887 reference library.

888

889 The three reference sequences used to supplement the reference library were:

890

891 >Lemmus_lemmus_16S_P007

892 TTAATTTTCCTGGCCTAACTTATAAACTTATACCTCTACTGAACTAAATAGTAAAGTCATAGG

893 CTAGCAATTTC

894 >Lemmus_sibiricus_16S_P007

895 TTAATTTTCCTGGCCTAACTTATAAACCTATACCTCTACTGAACTAAACAATAAAGTCATAGG

896 CTAGCAATTTC

897 >Microtus oeconomus_16S_P007_reconstructed

898 TTAATTTTCCTAGCTTAGCTACCCAACTTACTCCCCTACGGGACCAAAAAGAAAAGAAATAA

899 GCTAGTAATTTC

900

901

902

903

904

905

906

907 Supplementary references

908

- 909 1. Hughes, A. L. C., Gyllencreutz, R., Lohne, Ø. S., Mangerud, J. & Svendsen, J. I. The
910 last Eurasian ice sheets – a chronological database and time-slice reconstruction,
911 DATED-1. *Boreas* **45**, 1–45 (2016).
- 912 2. Patton, H. *et al.* Deglaciation of the Eurasian ice sheet complex. *Quat. Sci. Rev.* **169**,
913 148–172 (2017).
- 914 3. Pässe, T. & Andersson, L. Shore-level displacement in Fennoscandia calculated from
915 empirical data. *GFF* **127**, 253–268 (2005).
- 916 4. Corner, G. D., Yevzerov, V. Y., Kolka, V. V. & Møller, J. J. Isolation basin stratigraphy and
917 Holocene relative sea-level change at the Norwegian—Russian border north of Nikel,
918 northwest Russia. *Boreas* **28**, 146–166 (1999).
- 919 5. Stepanova, A. *et al.* Late Weichselian to Holocene history of the Baltic Sea as reflected
920 in ostracod assemblages. *Boreas* **48**, 761–778 (2019).
- 921 6. Bradley, S. L., Ely, J. C., Clark, C. D., Edwards, R. J. & Shennan, I. Reconstruction of
922 the palaeo-sea level of Britain and Ireland arising from empirical constraints of ice
923 extent: implications for regional sea level forecasts and North American ice sheet
924 volume. *J. Quat. Sci.* **38**, 791–805 (2023).
- 925 7. Giguët-Covex, C., Jelavić, S. & Foucher, A. The Sources and Fates of Lake
926 Sedimentary DNA. *Change Using Lake ...* (2023).
- 927 8. Ataman, T. G., Lammers, Y., Alsos, I. G., Rijal, D. P. & Brown, A. G. Sedimentary DNA
928 from lake depocenters maximizes detection of catchment vegetation. *Commun. Earth*
929 *Environ.* **6**, 1–10 (2025).
- 930 9. Brown, A. G. *et al.* New integrated molecular approaches for investigating lake
931 settlements in north-western Europe. *Antiquity* **96**, 1179–1199 (2022).

- 932 10. Picard, M. *et al.* Using DNA archived in lake sediments to reconstruct past ecosystems.
933 in *Encyclopedia of Quaternary Science* 673–690 (Elsevier, 2025).
- 934 11. Pedersen, M. W. *et al.* Ancient and modern environmental DNA. *Philosophical*
935 *Transactions of the Royal Society of London Series B-Biological Sciences* **370**,
936 20130383 (2015).
- 937 12. Sand, K. K., Jelavić, S., Kjær, K. H. & Prohaska, A. Importance of eDNA taphonomy and
938 sediment provenance for robust ecological inference: Insights from interfacial
939 geochemistry. *Environ. DNA* **6**, (2024).
- 940 13. Alvarez, A. J., Khanna, M., Toranzos, G. A. & Stotzky, G. Amplification of DNA bound on
941 clay minerals. *Mol. Ecol.* **7**, 775–778 (1998).
- 942 14. Kjær, K. H. *et al.* A 2-million-year-old ecosystem in Greenland uncovered by
943 environmental DNA. *Nature* **612**, 283–291 (2022).
- 944 15. Deiner, K., Fronhofer, E. A., Mächler, E., Walser, J.-C. & Altermatt, F. Environmental
945 DNA reveals that rivers are conveyor belts of biodiversity information. *Nat. Commun.* **7**,
946 12544 (2016).
- 947 16. Hardeng, J., Bakke, J., Sabatier, P., Støren, E. W. N. & Van der Bilt, W. Lake sediments
948 from southern Norway capture Holocene variations in flood seasonality. *Quat. Sci. Rev.*
949 **290**, 107643 (2022).
- 950 17. Morlock, M. A., Rodriguez-Martinez, S., Huang, D. Y.-T. & Klaminder, J. Erosion regime
951 controls sediment environmental DNA-based community reconstruction. *Environ. DNA*
952 **5**, 1393–1404 (2023).
- 953 18. Sjögren, P. *et al.* Lake sedimentary DNA accurately records 20th Century introductions
954 of exotic conifers in Scotland. *New Phytol.* **213**, 929–941 (2017).
- 955 19. Ficetola, G. F. *et al.* DNA from lake sediments reveals long-term ecosystem changes
956 after a biological invasion. *Sci. Adv.* **4**, eaar4292 (2018).
- 957 20. Giguët-Covex, C. *et al.* New insights on lake sediment DNA from the catchment:
958 importance of taphonomic and analytical issues on the record quality. *Sci. Rep.* **9**, 14676
959 (2019).

96021. Salonen, J. S. *et al.* Uncovering Holocene climate fluctuations and ancient conifer
961 populations: Insights from a high-resolution multi-proxy record from Northern Finland.
962 *Glob. Planet. Change* **237**, 104462 (2024).
96322. Rijal, D. P. *et al.* Sedimentary ancient DNA shows terrestrial plant richness continuously
964 increased over the Holocene in northern Fennoscandia. *Science Advances* **7**, eabf9557
965 (2021).
96623. Brown, A. G., Lucas, M., Alsos, I. G., Fromm, B. & Hudson, S. The sedaDNA revolution
967 and archaeology: Progress, challenges, and a research agenda. *J. Archaeol. Sci.* **174**,
968 106132 (2025).
96924. Capo, E. *et al.* Lake Sedimentary DNA Research on Past Terrestrial and Aquatic
970 Biodiversity: Overview and Recommendations. *Quaternary* **4**, 6 (2021).
97125. Liu, S. *et al.* Sedimentary ancient DNA reveals a threat of warming-induced alpine
972 habitat loss to Tibetan Plateau plant diversity. *Nat. Commun.* **12**, 2995 (2021).
97326. Liu, Y., Lisovski, S., Courtin, J., Stoof-Leichsenring, K. R. & Herzs Schuh, U. Plant
974 interactions associated with a directional shift in the richness range size relationship
975 during the Glacial-Holocene transition in the Arctic. *Nat. Commun.* **16**, 1128 (2025).
97627. Garcés-Pastor, S. *et al.* Wild and domesticated animal abundance is associated with
977 greater late-Holocene alpine plant diversity. *Nat. Commun.* **16**, 3924 (2025).
97828. Alsos, I. G. *et al.* Postglacial species arrival and diversity buildup of northern
979 ecosystems took millennia. *Science Advances* **8**, eabo7434 (2022).
98029. Nesje, A. A piston corer for lacustrine and marine sediments. *Arct. Alp. Res.* **24**,
981 257–259 (1992).
98230. Garcés-Pastor, S. *et al.* Terrestrial Plant DNA from Lake Sediments: Volume 6:
983 Sedimentary DNA. in *Tracking Environmental Change Using Lake Sediments* (eds.
984 Capo, E., Barouillet, C. & Smol, J. P.) vol. 21 275–298 (Springer International
985 Publishing, Cham, 2023).
98631. Leonard, J. A. *et al.* Animal DNA in PCR reagents plagues ancient DNA research. *J.*
987 *Archaeol. Sci.* **34**, 1361–1366 (2007).

98832. Wang, Y. *et al.* Late Quaternary dynamics of Arctic biota from ancient environmental
989 genomics. *Nature* (2021) doi:[10.1038/s41586-021-04016-x](https://doi.org/10.1038/s41586-021-04016-x).
99033. Renaud, G., Schubert, M., Sawyer, S. & Orlando, L. Authentication and assessment of
991 contamination in ancient DNA. *Methods Mol. Biol.* **1963**, 163–194 (2019).
99234. Heintzman, P. D. *et al.* The Sedimentary Ancient DNA Workflow. in *Tracking*
993 *Environmental Change Using Lake Sediments* (ed. Eric Capo, Cécilia Barouillet, John P.
994 Smol) 53–84 (Springer Nature, Cham, Switzerland, 2023).
99535. Schnell, I. B., Bohmann, K. & Gilbert, M. T. P. Tag jumps illuminated – reducing
996 sequence-to-sample misidentifications in metabarcoding studies. *Mol. Ecol. Resour.* **15**,
997 1289–1303 (2015).
99836. Valen, V., Mangerud, J., Larsen, E. & Hufthammer, A. K. Sedimentology and
999 stratigraphy in the cave Hamnsundhelleren, western Norway. *J. Quat. Sci.* **11**, 185–201
1000 (1996).
100137. Bachura, O. P. & Kosintsev, P. A. Holocene Large Mammal Fauna of the Central Part of
1002 the East European Plain. *Biology Bulletin* **47**, 996–1012 (2020).
100338. Salmi, A.-K., van den Berg, M., Niinimäki, S. & Pelletier, M. Earliest archaeological
1004 evidence for domesticated reindeer economy among the Sámi of Northeastern
1005 Fennoscandia AD 1300 onwards. *J. Anthropol. Archaeol.* **62**, 101303 (2021).
100639. Bråthen, K. A. *et al.* Endozoochory varies with ecological scale and context. *Ecography*
1007 **30**, 308–320 (2007).
100840. Steinhorsdottir, L. Investigating trophic interactions among herbivorous species in a
1009 rapidly changing Arctic tundra using DNA metabarcoding. (University of Oslo, 2019).
101041. Soininen, E. M. *et al.* Not only mosses: lemming winter diets as described by DNA
1011 metabarcoding. *Polar Biol.* **40**, 2097–2103 (2017).
101242. Soininen, E. M. *et al.* Shedding new light on the diet of Norwegian lemmings: DNA
1013 metabarcoding of stomach content. *Polar Biol.* **36**, 1069–1076 (2013).
101443. Unander, S., Mortensen, A. & Elvebakk, A. Seasonal changes in crop content of the
1015 Svalbard Ptarmigan *Lagopus mutus hyperboreus*. *Polar Res.* **3**, 239–245 (1985).

101644. Ingvaldsen, E. W. *et al.* Fecal DNA metabarcoding reveals seasonal and annual
1017 variation in willow ptarmigan diet. *R. Soc. Open Sci.* **11**, 231518 (2024).
101845. Garcés-Pastor, S. *et al.* High resolution ancient sedimentary DNA shows that alpine
1019 plant diversity is associated with human land use and climate change. *Nat. Commun.*
1020 **13**, 1–16 (2022).
102146. Barbero-Palacios, L. *et al.* Herbivore diversity effects on Arctic tundra ecosystems: a
1022 systematic review. *Environ. Evid.* **13**, 6 (2024).
102347. Kamenova, S. *et al.* Arctic greening drives changes in the diet and gut microbiome of a
1024 resident herbivore with consequences for fitness. *bioRxiv* 2024.02.27.582426 (2024)
1025 doi:[10.1101/2024.02.27.582426](https://doi.org/10.1101/2024.02.27.582426).
102648. Trondrud, L. M. *et al.* A summer heat wave reduced activity, heart rate, and autumn
1027 body mass in a cold-adapted ungulate. *Physiol. Biochem. Zool.* **96**, 282–293 (2023).
102849. Stark, S. *et al.* The ecosystem effects of reindeer (*Rangifer tarandus*) in northern
1029 Fennoscandia: Past, present and future. *Perspect. Plant Ecol. Evol. Syst.* **58**, 125716
1030 (2023).
103150. Ims, R. A. *et al.* Can reindeer overabundance cause a trophic cascade? *Ecosystems* **10**,
1032 607–622 (2007).
103351. Johnson, R. K. & Toprak, V. Local habitat is a strong determinant of spatial and temporal
1034 patterns of macrophyte diversity and composition in boreal lakes. *Freshw. Biol.* **66**,
1035 1490–1501 (2021).
103652. Brown, T. *et al.* Paleoeconomy more than demography determined prehistoric human
1037 impact in Arctic Norway. *PNAS Nexus* **1**, gac209 (2022).
103853. Giguët-Covex, C. *et al.* Long livestock farming history and human landscape shaping
1039 revealed by lake sediment DNA. *Nat. Commun.* **5**, (2014).
104054. Sievers, F. *et al.* Fast, scalable generation of high-quality protein multiple sequence
1041 alignments using Clustal Omega. *Mol. Syst. Biol.* **7**, 539 (2011).
- 1042 55. Gouy, M., Guindon, S. & Gascuel, O. SeaView version 4: A multiplatform graphical user
1043 interface for sequence alignment and phylogenetic tree building. *Mol. Biol. Evol.* **27**,

1044 221–224 (2010).

1045

